



*Supplement of*

## **Can biomonitors effectively detect airborne benzo[*a*]pyrene? An evaluation approach using modelling**

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## 22 Pine needles analysis and quantification

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24 Duplicate samples of 5 g of needles underwent ultrasonic extraction (USE) with a mixture of  
25 hexane:dichloromethane (1:1) as solvent and were subsequently cleaned-up using 5g alumina  
26 solid-phase extraction (SPE) cartridges from International Sorbent Technology (Mid  
27 Glamorgan, UK), using the same solvent for elution. After blowing down to dryness and  
28 solvent change to hexane, chromatographic analysis of BaP was done in a Varian CP-3800  
29 gas chromatograph (Lake Forest, CA, USA) coupled to a Varian 4000 mass spectrometer in  
30 Portugal and a Trace GC 2000 Series gas chromatograph from ThermoQuest (Waltham, MA,  
31 USA) coupled to a Finnigan Trace MS 2000 Series mass spectrometer in Spain. However,  
32 the operation was similar in both cases, namely using electron impact ionization (70 eV), a  
33 J&W Scientific (Folsom, CA, USA) 30 m × 0.25 mm I.D. DB-5 column coated with 5%  
34 diphenylpolydimethylsiloxane (film thickness 0.25 µm) and the same oven temperature  
35 program. The injector, transfer line and ion source temperatures were also the same (280, 250  
36 and 200 °C, respectively). Finally, the acquisition was made in single ion monitoring (SIM)  
37 mode using deuterated PAHs as surrogate standards. BaP was identified and quantified using  
38 retention time and up to three ions, with perylene-d<sub>12</sub> acting as surrogate standard and  
39 anthracene-d<sub>10</sub> as internal standard to look for GC-MS errors.

40 Linear behaviour between 0.01 and 1 mg L<sup>-1</sup> and good chromatographic resolution was  
41 obtained for BaP, with a limit of detection below 0.10 ng g<sup>-1</sup> (dry weight). The BaP  
42 concentrations were calculated in dry weight, after determining the water content of the  
43 needles for each species (Table S1). This information is needed for the estimates of air  
44 concentrations from the levels found in pine needles, as detailed below.

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46 Table S1. Characteristics of the four pine needle species employed in this study.

	<i>P. pinea</i>	<i>P. pinaster</i>	<i>P. halepensis</i>	<i>P. nigra</i>
<b>Mean mass of one needle (g)<sup>a</sup></b>	0.06	0.13	0.018	0.035
<b>Mean surface area (m<sup>2</sup>×10<sup>-6</sup>)<sup>a</sup></b>	545	815	254	366
<b>Lipid content (mg g<sup>-1</sup>, dw)</b>	121.95	182.93	105.56	104.26
<b>Water content (% mass)</b>	59	59	46	53

47 <sup>a</sup> Data taken from Daligault (1991) and Moro (2006)

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49 **Modelling experiment**

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51 Table S2. Set of parameterisations used in the WRF+CHIMERE modelling system

WRF	CHIMERE
Microphysics → <b>WSM3</b>	Chemical Mechanisms → <b>MELCHIOR2</b>
PBL → <b>Yonsei University</b>	Aerosol chemistry → Inorganic (thermodynamic equilibrium
Radiation → <b>CAM</b>	with <b>ISORROPIA</b> ) and organic ( <b>MEGAN SOA</b> scheme)
Soil → <b>Noah LSM</b>	aerosol chemistry
Cumulus → <b>Kain-Fritsch</b>	Natural aerosols → <b>dust, re-suspension and inert sea-salt</b>
	<b>BC → LMDz-INCA+GOCART</b>

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53 The Advanced Research Weather Research and Forecasting (WRF-ARW) Model v3.1.1  
 54 (Klemp et al., 2007; Skamarock et al., 2008) is used to provide the meteorology to the  
 55 chemistry transport models. WRF is a fully compressible, Eulerian non-hydrostatic model  
 56 that solves the equations that govern the atmospheric motions. 33 vertical layers on sigma  
 57 coordinates cover from the ground level up to 10 hPa. Microphysical processes are treated  
 58 using the single-moment 3-class scheme described in Hong et al. (2004). The sub-grid-scale  
 59 effects of convective and shallow clouds are resolved by a modified version of the Kain-  
 60 Fritsch scheme based on Kain and Fritsch (1993). The Noah land surface model was used to  
 61 solve the soil processes on 4 layers to a depth of 2m (Chen and Dudhia, 2001a; 2001b). The  
 62 vertical sub-grid-scale fluxes caused by eddy transport in the atmospheric column are  
 63 resolved by the Yonsei University non-local planetary boundary layer scheme (Noh et al.,  
 64 2003). Finally, radiation was treated through the Community Atmospheric Model (CAM) 3.0  
 65 radiation scheme (Collins et al., 2006).

66 WRF was coupled off-line to CHIMERE. Atmospheric concentrations of BaP have been  
 67 calculated using CHIMERE chemistry transport model (v2008b), coupled off-line to WRF  
 68 outputs and EMEP emissions. For further details on the model options, the reader is referred  
 69 to Menut et al. (2013). MELCHIOR2 gas-phase mechanism is implemented within  
 70 CHIMERE. The chemistry transport model includes aerosol and heterogeneous chemistry;  
 71 distinguishes among different chemical aerosol components, namely nitrate, sulphate,

72 ammonium, elemental and organic carbon with three subcomponents (primary, secondary  
73 anthropogenic and secondary biogenic) and marine aerosols. Unspecified primary  
74 anthropogenic aerosols and aerosol water are additionally kept as separate components. The  
75 model considers the thermodynamic equilibrium using the ISORROPIA model (Nenes et al.,  
76 1998). Last, the aerosol microphysical description for CHIMERE is based on a sectional  
77 aerosol module including 6 bins from 10 nm to 40  $\mu\text{m}$  using a geometrical progression.  
78 Moreover, a dynamical approach is used to describe the gas/particle conversion, in line with  
79 Bowman et al. (1997):

$$80 \quad J_i = 1/\tau_i (G_i - G_{i_{eq}})$$

81 Where  $J_i$  ( $\mu\text{g m}^{-3} \text{s}^{-1}$ ) is the absorption or desorption flux of species  $i$ ;  $\tau_i$  (s) is a characteristic  
82 time of the mass transfer that is a function of particle size and the chemical properties of  $i$ ;  $G_i$   
83 is the bulk gas-phase concentration of  $i$  and  $G_{i_{eq}}$  is the gas-phase concentration of  $i$  at  
84 equilibrium. The gas-phase concentrations at equilibrium depend on the chemical  
85 composition of the particles, the temperature and, for hydrophilic species, the relative  
86 humidity (Pun et al., 2006).

87 In the present work, simulations covered the period 2006-2010. Initial and boundary  
88 conditions for WRF were provided by ERA-Interim reanalysis (Dee et al., 2011), while for  
89 CHIMERE, the global climate chemistry model LMDz-INCA2 was used (96 x 72 grid cells,  
90 namely 3.75° x 2.5° in longitude and latitude, with 19 sigma-p hybrid vertical levels, Szopa et  
91 al. (2009) developed by the Laboratoire des Sciences du Climat et l'Environnement (LSCE).  
92 Climatic monthly mean data are interpolated in the horizontal and vertical dimensions to  
93 force the major chemical concentrations at the boundaries of the domain. A detailed  
94 description of the INteractive Chemistry and Aerosol (INCA) model is presented by  
95 Hauglustaine et al. (2004) and Folberth et al. (2006). Because the contribution of long-range  
96 transport on ground level concentrations (those considered in this work) can be considered as  
97 negligible, the influence of using climatological boundary conditions is limited and  
98 overwhelmed by local processes.

99 Anthropogenic emissions for the entire period of simulations are derived from the EMEP  
100 database (Vestreng et al., 2009) and disaggregated to the working resolution following spatial  
101 proxy data, according to the methodology stated in Pay et al. (2010). For BaP emissions, data  
102 have been obtained from the EMEP-MSCEAST web site (<http://www.msceast.org>). The  
103 accuracy of simulations depends strongly on emission data and unfortunately there are strong

104 uncertainties in BaP emissions, by a factor of 2 to 5 (San José et al., 2013). According to  
105 these authors, the main source of BaP is incomplete combustion processes of organic  
106 material, in particular wood and coal in private households. Industrial heating and cookeries  
107 as well as road traffic are also large sources of BaP, which is emitted in particle phase.

108 Natural emissions (of sea salt and dust) depend on meteorological conditions, and  
109 consequently they are coupled hourly to WRF meteorological outputs. Biogenic emissions  
110 were generated dynamically using MEGAN (Model of Emissions of Gases and Aerosols  
111 from Nature) (Guenther et al., 2006) with the parameterized form of the canopy environment  
112 model. The model estimates hourly isoprene, monoterpene, and other BVOC emissions based  
113 on plant functional type and as a function of hourly temperature and ground level shortwave  
114 radiation from WRF.

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## 116 **Model validation**

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118 EMEP stations are located at a minimum distance of approximately 10 km from large  
119 emission sources and thus assumed to fit the resolution of the model used for regional  
120 background concentrations (Torseth et al., 2012). Thus, as reported by Ratola and Jiménez-  
121 Guerrero (2015), results from the EMEP monitoring data were used to characterize the ability  
122 of the model to reproduce present air BaP levels and variability. The “EMEP Manual for  
123 Sampling and Analysis” (EMEP, 2001) describes all the sampling methodologies employed  
124 for each chemical and/or matrix and the recommended operation, as well as the data quality  
125 objectives for the yielded results. Final The available stations running in the Iberian  
126 Peninsula in the 2006-2010 time frame were: Niembro (2006-2010), Campisabalos (2007-  
127 2008), O Saviñao (2007), Víznar (2008-2010), Peñausende (2008-2009), Barcarrota (2008),  
128 Zarra (2008), San Pablo de los Montes (2009-2010), Mahón (2010) and Els Torms (2010). In  
129 all of them, BaP frequencies of measurement and duration varied probably depending on the  
130 budget limitations, but when sampling campaigns were active, they were performed usually  
131 once a week. The handling of samples is taken with extreme care to limit external  
132 contaminations and/or degradation reactions to occur. For the more volatile chemicals, there  
133 is a bigger risk of having some losses, but in the case of BaP, since it is almost all formed by  
134 particulate matter, it is bound to stay stable under the appropriate storage conditions  
135 (commonly in the freezer until analysis). The results (available as weekly or monthly

136 averages) were compared to the available periods for observations. Regarding the  
137 uncertainty, no information is given for the Iberian sites, but it generally should meet the  
138 EMEP data quality objectives for the combined sampling and chemical analysis (between 15  
139 and 25%) (EMEP, 2001).

140 Being well aware of the need for further measurements with a higher temporal coverage, the  
141 strong limitation (not only over the Iberian Peninsula, but worldwide) for simultaneous air  
142 and vegetation measurements forced us to rely on the best information available. In doing so,  
143 this work intends to set a starting point for an improvement in the design of sampling  
144 campaigns and associated modelling strategies. Although it was possible to find some data  
145 from air monitoring stations from the Generalitat de Catalunya and the Comunitat  
146 Valenciana, not all of them presented climatologically representative series. Thus, also to  
147 maintain a wider geographical coverage with under the same sampling and analytical  
148 framework to ensure the homogeneity of the data.

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150 For the evaluation of canopy deposition and atmospheric concentrations, a number of  
151 statistical parameters have been selected (Figure S1). Spatial correlation coefficient ( $r$ ), root  
152 mean square error (RMSE) and mean bias (MB) values are commonly used by the modelling  
153 community and have therefore been selected according to the criteria of Pay et al. (2010),  
154 who use them to evaluate a modelling system for Europe (“bias” is intended as the difference  
155 between modelled and observed means). Moreover, Boylan and Russell (2006) suggest that  
156 the mean normalised bias error (MNBE) for each model-observed pair by the observation is a  
157 useful parameter, but may not be appropriate for evaluating particulate matter and their  
158 components. These authors suggested the mean fractional bias (MFB) and the mean  
159 fractional error (MFE) instead, indicating that model performance goal would be met when  
160 both the MFE and MFB are less than or equal to 50% and  $\pm 30\%$ , respectively, and the model  
161 performance criterion when  $MFE \leq 75\%$  and  $MFB \leq \pm 60\%$ . These criteria and goals have  
162 been selected to provide the metrics for the WRF+EMEP+CHIMERE evaluation of BaP.  
163 Annual and seasonal mean statistics are computed, with seasons corresponding to December,  
164 January and February (DJF, winter), March, April and May (MAM, spring), June, July and  
165 August (JJA, summer) and September, October and November (SON, autumn).

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MOD MEAN (modelled concentrations)	$\frac{1}{N} \sum C_{mod}$
OBS MEAN (pine needle concentrations)	$\frac{1}{N} \sum C_{obs}$
BIAS	$\frac{1}{N} \sum (C_{mod} - C_{obs})$
RMSE (root mean square error)	$\sqrt{\sum \frac{(C_{mod} - C_{obs})^2}{N}}$
MFB (mean fractional bias)	$\frac{1}{N} \sum \left( \frac{(C_{mod} - C_{obs})}{\left(\frac{C_{mod} + C_{obs}}{2}\right)} \right)$

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Figure S1. Main statistical parameters used in model validation

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As our aim is to have the best approximation of atmospheric BaP levels through modelling procedures, to serve as a reference pseudo-reality to estimate the most accurate vegetation-to-air conversion method, the multiplicative ratio bias-adjustment technique has been applied following the methodology of Borrego et al. (2011). The adjustment factor is calculated as the quotient between the additions of observed and modelled concentrations at a particular hour of the  $n$  previous days. Borrego et al. (2011) and Monteiro et al. (2013) recommend a four-day training period ( $n=4$ ). However, given the limited availability of EMEP data (only on a weekly basis), a four-week training period has been chosen here instead as a compromise between having a sufficiently long timeframe to gather adequate statistics but not as much as to mask seasonal variations. This bias-adjustment technique improves the relative mean bias (expressed as percentage) by approximately 90% (Monteiro et al., 2013). However, the goal is to remove potential systematic model errors intrinsic to each model formulation or input data, rather than obtaining an additional assessment of the possible model flaws or performance or to correct them artificially. Figure S2 depicts the mathematical representation of this approach, with  $C^{\text{corrected}}$ ,  $C^{\text{model}}$ , and  $C^{\text{obs}}$  as the bias-adjusted, original modelled and measured concentrations at a given hour “h” and day “day”.

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$$C^{\text{corrected}}(h, \text{day}) = -\frac{1}{n_{\text{days}}} \sum_{n_{\text{days}}} (C_h^{\text{model}} - C_h^{\text{obs}}) + C^{\text{model}}(h, \text{day})$$

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Figure S2. Mathematical expression for the bias-adjustment of the modelled results.

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As stated in Monteiro et al. (2013), the global mean bias is minimised for all the monitoring stations, using the bias detected in previous days for a given hour (h) of the day. These procedures are model, site, and time of day specific.

191 **Results**

192 Table S3. Parameters of the modelled deposition over vegetal canopies evaluated against  
 193 observations compiled from pine needles, for all the sampling points (n – number of  
 194 duplicate samples; mean concentrations in ng g<sup>-1</sup>).

SITE	n	LAT	LONG	PINE SPECIES	BIAS	MFB	OBS. MEAN	MOD. MEAN
Alcolea de Cinca	1	42.03	-1.56	<i>Pinus pinea</i>	-0.63	-95.41%	0.98	0.35
Alcoutim	4	37.47	-7.47	<i>Pinus pinea</i>	0.11	26.63%	0.81	0.92
Antuã 1	4	40.69	-8.52	<i>Pinus pinea</i>	-0.17	24.18%	2.71	2.53
Barcelona	1	41.39	2.11	<i>Pinus pinea</i>	-2.53	-105.46%	3.66	1.13
Beja	4	38.01	-7.87	<i>Pinus pinea</i>	-0.29	20.86%	1.02	0.73
Braga	4	41.56	-8.40	<i>Pinus pinea</i>	0.71	31.72%	0.96	1.67
Castelo Branco	4	39.83	-7.50	<i>Pinus pinea</i>	0.60	31.72%	0.81	1.41
Coimbra	4	40.21	-8.42	<i>Pinus pinea</i>	0.54	32.59%	0.62	1.16
El Bocal	1	41.57	-0.69	<i>Pinus pinea</i>	-0.49	-33.85%	1.71	1.21
El Prat	1	41.30	2.10	<i>Pinus pinea</i>	-0.38	-16.77%	2.44	2.06
Évora	4	38.58	-7.91	<i>Pinus pinea</i>	-1.13	6.74%	1.33	0.21
Faro	4	37.02	-7.94	<i>Pinus pinea</i>	-1.53	7.34%	1.85	0.32
Leiria	4	39.75	-8.80	<i>Pinus pinea</i>	0.34	29.56%	0.76	1.10
Lisboa	4	38.72	-9.14	<i>Pinus pinea</i>	-4.73	5.32%	5.37	0.64
Loulé	4	37.13	-8.10	<i>Pinus pinea</i>	-1.90	10.17%	2.56	0.65
Maleján	1	41.82	-1.55	<i>Pinus pinea</i>	-0.77	-91.95%	1.22	0.45
Miranda de Ebro 1	1	42.68	-2.95	<i>Pinus pinea</i>	-0.25	-70.21%	0.49	0.23
Monteagudo	1	41.96	-1.69	<i>Pinus pinea</i>	-0.34	-26.47%	1.46	1.12
Movera	1	41.64	-0.80	<i>Pinus pinea</i>	-0.01	-0.61%	1.22	1.21
Outão	4	38.49	-8.98	<i>Pinus pinea</i>	2.11	35.21%	1.53	3.64
Portalegre	4	39.30	-7.43	<i>Pinus pinea</i>	-0.01	24.89%	1.24	1.23
Porto 1	4	41.18	-8.60	<i>Pinus pinea</i>	1.08	31.13%	1.66	2.74
Praia Verde	4	37.18	-7.48	<i>Pinus pinea</i>	-0.22	17.50%	0.47	0.25
Quintãs 1	4	40.58	-8.63	<i>Pinus pinea</i>	0.80	33.80%	0.74	1.53
Santarém	4	39.24	-8.69	<i>Pinus pinea</i>	-0.73	16.55%	1.44	0.71
Sines	4	37.96	-8.81	<i>Pinus pinea</i>	0.03	25.51%	0.75	0.78
Souselas	4	40.29	-8.41	<i>Pinus pinea</i>	1.58	29.94%	3.20	4.78
Torres de Segre	1	41.54	0.51	<i>Pinus pinea</i>	-0.11	-7.74%	1.46	1.35
Vic	1	41.94	2.25	<i>Pinus pinea</i>	-0.71	-21.37%	3.66	2.95
Villodas	1	42.83	-2.78	<i>Pinus pinea</i>	1.91	98.82%	0.98	2.88
Antuã 2	4	40.69	-8.52	<i>Pinus pinaster</i>	-0.67	22.50%	3.71	3.03
Bragança	4	41.81	-6.76	<i>Pinus pinaster</i>	0.23	26.96%	1.37	1.60
Caminha	4	41.87	-8.86	<i>Pinus pinaster</i>	0.54	29.23%	1.33	1.87
Estarreja	4	40.77	-8.57	<i>Pinus pinaster</i>	1.34	31.68%	1.83	3.17
Fóia	4	37.31	-8.61	<i>Pinus pinaster</i>	0.84	35.29%	0.60	1.44
Guarda	4	40.54	-7.27	<i>Pinus pinaster</i>	0.66	29.41%	1.55	2.21
Leça	4	41.22	-8.71	<i>Pinus pinaster</i>	-0.63	23.80%	6.85	6.22
Mirandela	4	41.37	-7.14	<i>Pinus pinaster</i>	-1.14	18.88%	2.89	1.76
Porto 2	1	41.18	-8.60	<i>Pinus pinaster</i>	1.20	28.27%	3.66	4.86
Quintãs 2	4	40.58	-8.63	<i>Pinus pinaster</i>	-0.14	24.13%	2.07	1.93
Rio de Onor	4	41.94	-6.61	<i>Pinus pinaster</i>	0.73	31.06%	1.14	1.87
Torre	4	40.31	-7.58	<i>Pinus pinaster</i>	0.32	29.64%	0.71	1.03
Vide	1	40.29	-7.78	<i>Pinus pinaster</i>	1.19	65.60%	1.22	2.41
Vila Real	4	41.30	-7.74	<i>Pinus pinaster</i>	2.17	32.42%	2.57	4.74
Arazuri	1	42.81	-1.72	<i>Pinus nigra</i>	0.14	20.40%	0.64	0.78
Briñas	1	42.59	-2.84	<i>Pinus nigra</i>	1.30	75.67%	1.06	2.36
La Bordeta	1	41.60	0.62	<i>Pinus nigra</i>	-0.32	-117.75%	0.43	0.11
Miranda de Ebro 2	1	42.67	-2.09	<i>Pinus nigra</i>	-0.10	-27.59%	0.43	0.32
Nestares	1	43.00	-4.15	<i>Pinus nigra</i>	0.00	-0.10%	0.43	0.43
Urdiáin	1	42.90	-2.14	<i>Pinus nigra</i>	0.61	83.80%	0.43	1.04
Amposta	1	40.72	0.58	<i>Pinus halepensis</i>	-0.60	-39.43%	1.83	1.23
Andosilla	1	42.37	-1.94	<i>Pinus halepensis</i>	0.38	29.17%	1.10	1.48
Caldearenas	1	42.40	-0.50	<i>Pinus halepensis</i>	0.01	3.14%	0.37	0.38
Cascante	1	41.98	-1.68	<i>Pinus halepensis</i>	-0.44	-63.19%	0.92	0.48
Cuarte de Huerva	1	41.61	-0.92	<i>Pinus halepensis</i>	-0.33	-21.98%	1.65	1.32



195 Table S3. (cont.) Parameters of the modelled deposition over vegetal canopies evaluated  
 196 against observations compiled from pine needles, for all the sampling points (n – number of  
 197 duplicate samples; mean concentrations in ng g<sup>-1</sup>).  
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<b>SITE</b>	<b>n</b>	<b>LAT</b>	<b>LONG</b>	<b>PINE SPECIES</b>	<b>BIAS</b>	<b>MFB</b>	<b>OBS. MEAN</b>	<b>MOD. MEAN</b>
Deltebre	1	40.71	0.71	<i>Pinus halepensis</i>	<b>-0.58</b>	<b>-37.70%</b>	<b>1.83</b>	<b>1.25</b>
Estella/Lizarra	1	42.67	-2.03	<i>Pinus halepensis</i>	<b>1.40</b>	<b>97.49%</b>	<b>0.73</b>	<b>2.13</b>
Flix	1	41.23	0.55	<i>Pinus halepensis</i>	<b>0.07</b>	<b>11.87%</b>	<b>0.55</b>	<b>0.62</b>
Grisén	1	41.73	-1.18	<i>Pinus halepensis</i>	<b>-1.22</b>	<b>-39.69%</b>	<b>3.67</b>	<b>2.45</b>
Logroño 1	1	42.47	-2.44	<i>Pinus halepensis</i>	<b>-0.44</b>	<b>-35.19%</b>	<b>1.47</b>	<b>1.03</b>
Logroño 2	1	42.67	-2.42	<i>Pinus halepensis</i>	<b>1.60</b>	<b>34.34%</b>	<b>3.85</b>	<b>5.45</b>
Mollerussa	1	41.62	0.91	<i>Pinus halepensis</i>	<b>-0.72</b>	<b>-77.74%</b>	<b>1.28</b>	<b>0.57</b>
Puente La Reina	1	42.67	-1.82	<i>Pinus halepensis</i>	<b>0.79</b>	<b>60.19%</b>	<b>0.92</b>	<b>1.71</b>
San Adrián	1	42.33	-1.93	<i>Pinus halepensis</i>	<b>-0.11</b>	<b>-8.91%</b>	<b>1.28</b>	<b>1.17</b>
Sástago	1	41.32	-0.34	<i>Pinus halepensis</i>	<b>-0.39</b>	<b>-72.55%</b>	<b>0.73</b>	<b>0.34</b>
Tornabous	1	41.69	1.05	<i>Pinus halepensis</i>	<b>-0.41</b>	<b>-77.45%</b>	<b>0.73</b>	<b>0.32</b>
Tortosa	1	40.80	0.51	<i>Pinus halepensis</i>	<b>-0.30</b>	<b>-16.01%</b>	<b>2.02</b>	<b>1.72</b>
Tudela 1	1	42.07	-1.60	<i>Pinus halepensis</i>	<b>-0.89</b>	<b>-35.90%</b>	<b>2.94</b>	<b>2.04</b>
Tudela 2	1	42.08	-1.62	<i>Pinus halepensis</i>	<b>-0.41</b>	<b>-25.29%</b>	<b>1.83</b>	<b>1.42</b>
Villanueva de Gállego	1	41.77	-0.82	<i>Pinus halepensis</i>	<b>-0.74</b>	<b>-31.16%</b>	<b>2.75</b>	<b>2.01</b>

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