

Supplement of Atmos. Chem. Phys., 15, 1221–1236, 2015
<http://www.atmos-chem-phys.net/15/1221/2015/>
doi:10.5194/acp-15-1221-2015-supplement
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Supplement of

Can positive matrix factorization help to understand patterns of organic trace gases at the continental Global Atmosphere Watch site Hohenpeissenberg?

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9 **Supplement**

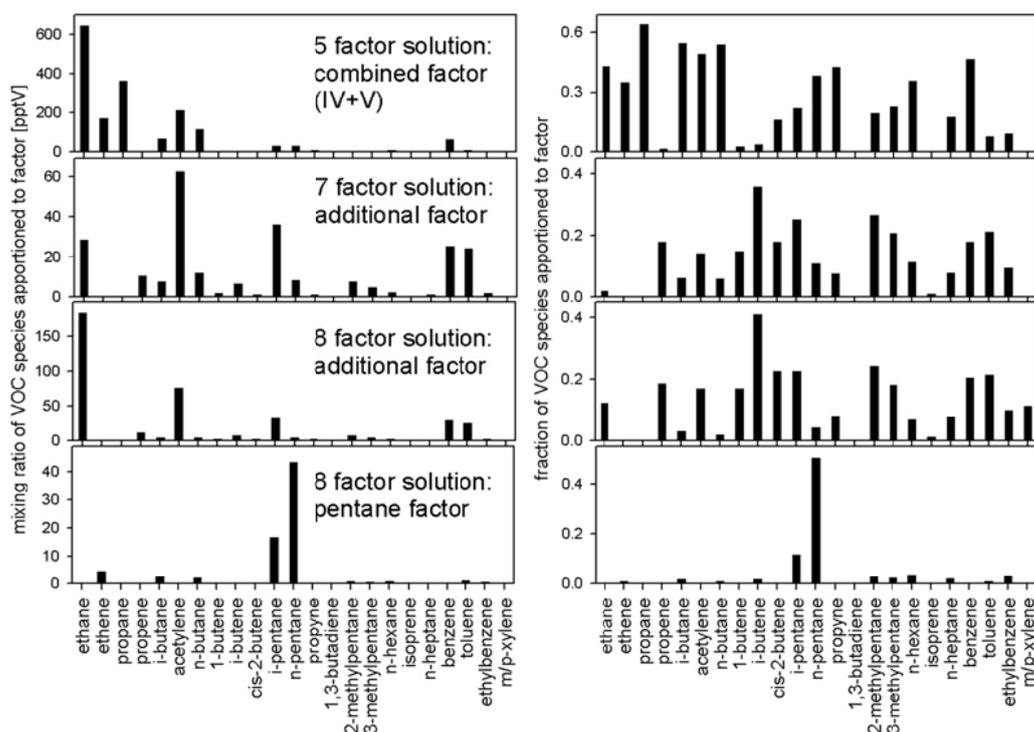
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11 **S1. Comparison to alternative factor solutions**

12 In Fig. S1 the additional factors that were resolved in the seven and eight factor
13 solutions and one different factor for the five factor solution are shown. In
14 general, all factors of the six factor solution (Fig. 3) could also be found in these
15 alternative computations. For the five factor computation considerable differences
16 to the solution above could be found in Factors II, IV, and V. Factor II also
17 contained large amounts of ethane, smaller amounts of acetylene, benzene, and
18 toluene. The longer-lived Factors IV and V were combined to one factor. The
19 seven and eight factor solutions provided an additional common factor that
20 contained more than one third of the modeled isobutene and considerable amounts
21 of 2-methylpentane, 3-methylpentane, benzene, toluene and some other
22 substances. Thus, the amounts of these substances in other factors, especially
23 Factor III, decreased accordingly, compared to the six factor computation. Ethane
24 in Factor V only occurred for the six factor solution; it was distributed in other
25 factors, e.g. Factor VI, for the seven and eight factor solutions. Isopentane and *n*-
26 pentane accounted for the second additional factor in the eight factor solution,
27 thus, their contribution to Factor III was much lower than in the six factor
28 solution.

29 The five factor solution combined the two long-lived Factors IV and V from the
30 six factor solution that should appear separately when compared e.g. to Sauvage
31 et al. (2009) as well as to the source apportionment in section 3.2. The seven and
32 eight factor solutions showed an additional factor with comparatively small
33 contributions of many substances that could not be apportioned to proper source

37 categories. Since the statistical values of the PMF analysis did not show a clear
 38 indication for the ideal number of factors, the decision was made to choose six
 39 factors as the most reasonable and explainable solution that PMF calculated.
 38



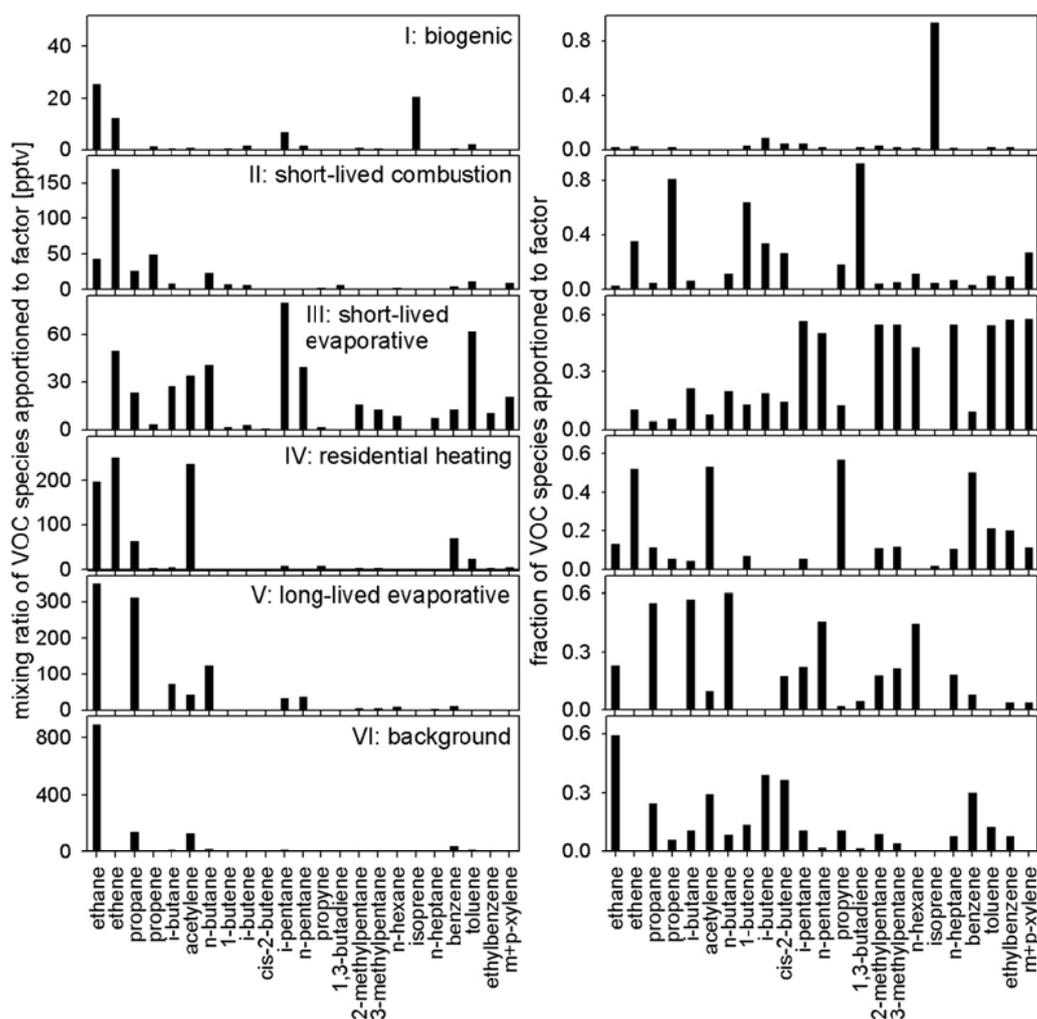
39
 42 Figure S1. Factor profiles of alternative solutions calculated by PMF, left: mixing
 43 ratio of each species apportioned to each factor [pptv], right: contribution of each
 44 factor to the species.

43

44 S2. Resolved factors including nighttime data

54 Figure S2 shows factor profiles for the six factor solution calculated by PMF for
 55 all measured data (daytime and nighttime). The same six factors (compared to
 56 Fig. 3) were resolved and only slight differences to the daytime data solution were
 57 found which supports the stability of the PMF solution in extracting profiles
 58 reliably also at this remote site. One main difference could be seen for the
 59 biogenic factor. The relative contributions remained basically identical, but the
 60 absolute amount was lower by a factor of around 2. This can be explained by very
 61 low to zero emissions of isoprene, the dominant compound, at night. The five
 62 anthropogenic factors were also basically identical. Only small additional
 63 abundances of ethane and propane as well as some higher contribution of the

55 dominating ethene in the short-lived combustion factor could be observed.



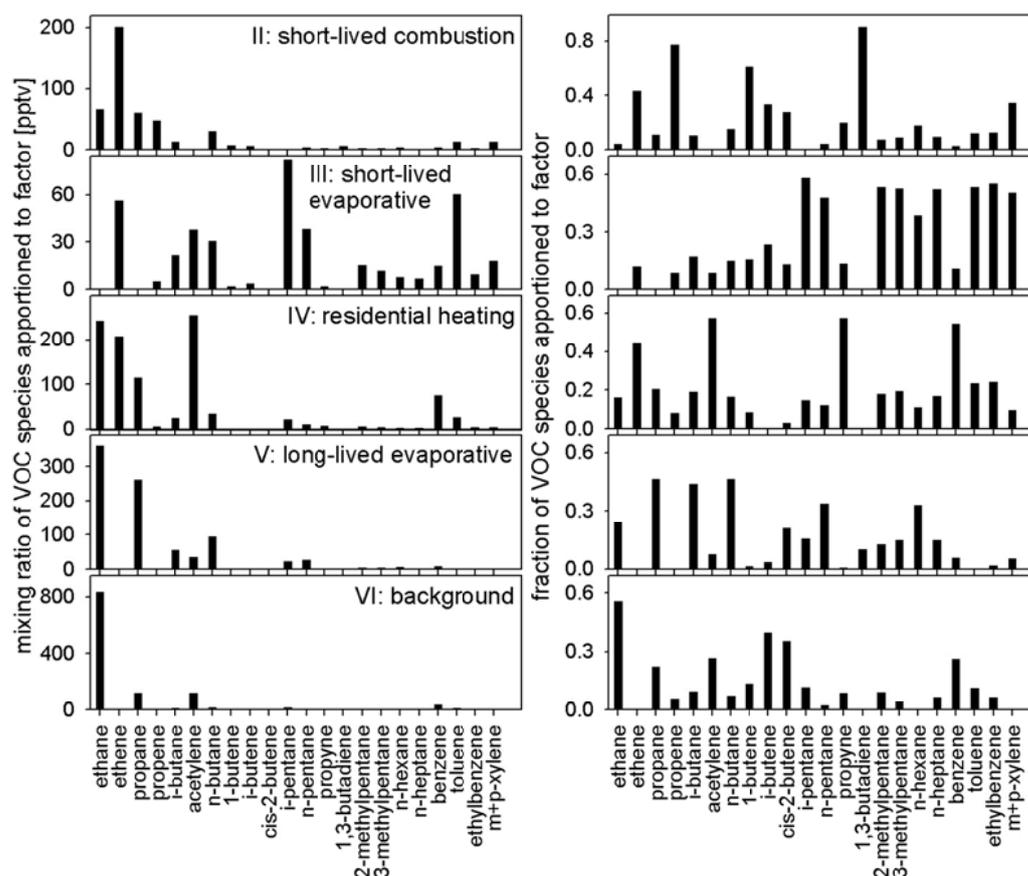
56
 60 Figure S2. Factor profiles of the six factor solution calculated by PMF for all data
 61 (daytime and nighttime), left: mixing ratio of each species apporntioned to each
 62 factor [pptv], right: contribution of each factor to the species. Note that the scales
 63 for each subplot are different due to large variations in absolute mixing ratios.

61

62 S3. Resolved factors from nighttime data only

70 In Fig. S3 the PMF results from exclusively nighttime data are plotted. Only five
 71 reasonable factors were extracted, because the biogenic factor disappeared due to
 72 low or no isoprene contributions at nighttime. However, all five anthropogenic
 73 factors were resolved and exhibited quite similar profiles as for daytime (Fig. 3)
 74 or daytime and nighttime (Fig. S2) data, respectively. Small differences could be
 75 found in the short-lived combustion factor with higher contribution of ethene and
 76 some small additional contribution of propane. The residential heating/long-lived
 77 combustion factor showed higher values of ethane and acetylene and slightly

71 lower values of ethene. Ethane was higher in the long-lived evaporative factor.



72

78 Figure S3. Factor profiles of the five factor solution calculated by PMF for
79 nighttime data only, left: mixing ratio of each species apportioned to each factor
80 [pptv], right: contribution of each factor to the species. Note that the scales for
81 each subplot are different due to large variations in absolute mixing ratios. At
82 nighttime only five factors were resolved, since isoprene contributions were too
83 low for a separate factor.

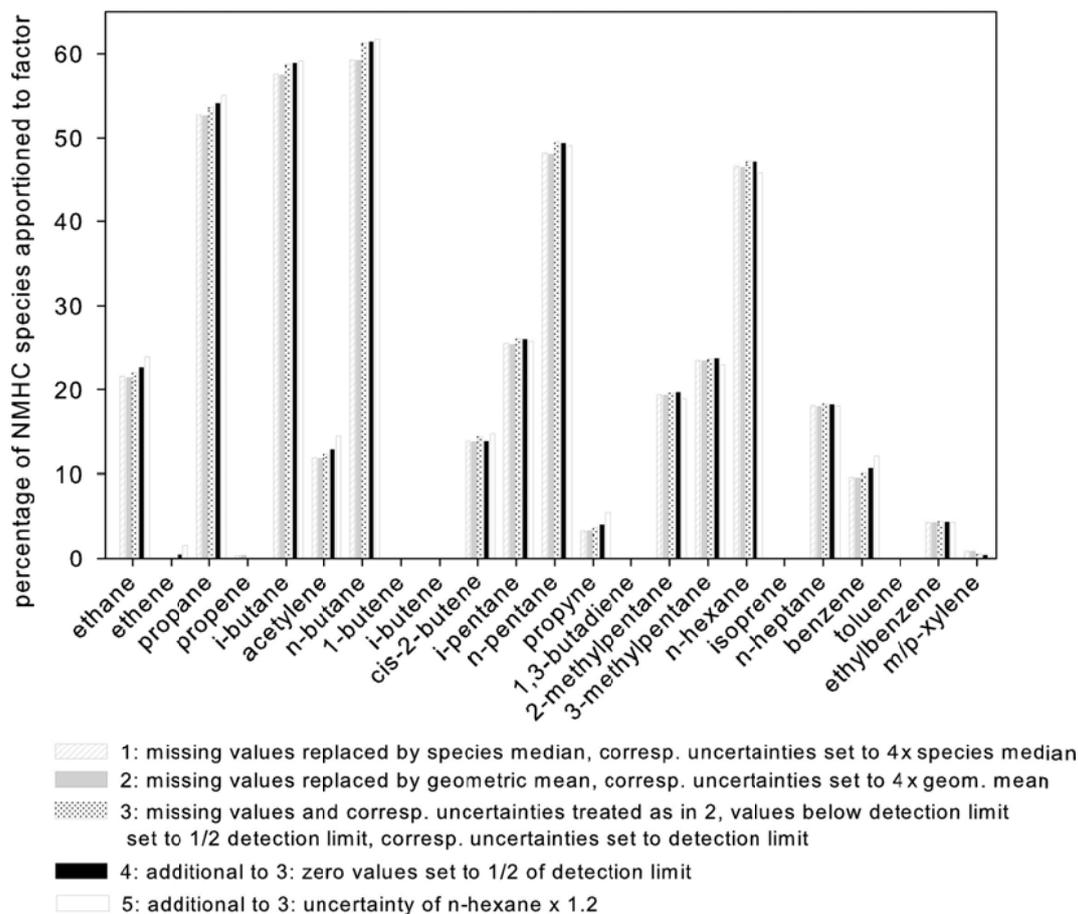
79

80 **S4. Effects of different data treatment methods**

87 Five methods to deal with missing values, values below the detection limit and
88 uncertainty values due to interference with other substances or other measurement
89 problems were applied to the six factor solution to analyze the effects of different
90 treatments to the overall solution (cf. Tab. 1). Figure S4 shows the results of the
91 five different scenarios displayed on a relative scale. Only the results for Factor V
92 are shown, since within this factor the largest differences between the computed
93 results of different datasets can be found.

88 No significant statistical difference for any profile was found for any of the five

93 treatments with a Levene's test for variances. Thus, in this study different
 94 treatment of these values had only a very small quantitative impact on the final
 95 result. An additional reason for the differences between the computations could be
 96 different starting points in PMF that led to slightly different resulting numbers,
 97 but had no influence on the overall stable result.
 98



95
 98 Figure S4. Composition profiles of Factor V, showing the relative contribution of
 99 each substance to the factor, for five different treatments of missing and small
 100 values (see Tab. 1).
 99

101 **S5. Inclusion of a rough proxy for profile distortion due to**
 102 **photochemical aging**

105 Based on the rate constants for the respective substances (Atkinson, 2000), a
 106 reactivity error value depending on different transport or reaction time was
 107 calculated for every sample and every species (cf. section 2.3) following the
 108 approach by Sauvage et al. (2009). Reactivity error contributions were calculated

105 for up to 72 hours. However, it should be pointed out that the variation of the
106 transport times around characteristic values should be considered in relation to the
107 respective factor and the lifetimes of compounds present in the factor.
108 Accordingly, for Factors II and III with compound lifetimes of hours to few days,
109 it does not make sense to vary the transport times by more than 12-24 hours,
110 otherwise compounds vanish. Furthermore, it should be considered that a factor is
111 distorted due to aging depending on the different lifetimes of the VOC
112 compounds present in the respective factor. But the factors determined by PMF in
113 this paper have been demonstrated to represent compounds with similar lifetimes.
114 Thus, the distorting effect is rather small, and in fact the uncertainty due to aging
115 should be scaled to lifetime differences rather than absolute lifetimes. However,
116 we want to use and discuss the Sauvage et al. (2009) approach, and Tab. S1 shows
117 the average values of the calculated uncertainties for modeled reaction times of 0
118 and 36 hours exemplarily. Uncertainties for compounds with low reactivity such
119 as some alkanes and aromatic species remained at relatively low levels; ethane as
120 the most stable substance changed its uncertainty from 7.1 to 8.0%, benzene from
121 8.6 to 16.3%. A stronger influence of reactivity existed for alkenes, e.g. for 1-
122 butene from 29.3 to 92.4% or 1,3-butadiene from 43.6 to 108.9%. The uncertainty
123 for the highly reactive isoprene rose from 33.3% to 101.7% after 12 h (not shown)
124 and to 115.8% after 36 h of assumed reaction time. These high uncertainties made
125 proper modeling in PMF impossible and for the reasons discussed before this
126 demonstrates the limits of this approach. To partly overcome this, the isoprene
127 uncertainty due to reactivity was not considered as the biogenic factor mainly
128 depends on isoprene alone.

129 The attempt to compensate the mass conservation assumption of the PMF model
130 by changing the uncertainty values of the reactive species had an impact on the
131 resulting factors. The model reacted to the change of uncertainty data by changing
132 the quantitative composition of each factor; the Q values for the model decreased
133 with increasing uncertainties. The largest change of substance contributions to
134 each factor was found between zero and twelve hours where the uncertainties
135 changed the most. Despite the sometimes large changes of some substance
136 contributions, especially ethane and ethene, the relative profiles of the factors
137 remained more or less recognizable for the different tested times. This
138 demonstrates the robustness of the results even though this approach is rather

139 crude for the reasons discussed above. The absolute and relative factor profiles of
140 the PMF solution with six factors and 36 h reaction time for all substances except
141 isoprene are shown in Fig. S5 to illustrate the changes in the solutions with
142 integrated reactivity. However, due to the discussion above the apparently small
143 deviations to the original solution without introducing this enhanced uncertainty
144 due to reactivity were not further discussed.

145 In terms of model stability, bootstrap runs for the 0 hours solution were very
146 stable with only small variations in the contributions of the substances to the
147 factors. Small variability could only be seen for Factors II and III in regards to
148 some alkenes. Bootstrap runs including reaction times of 12-72 hours were still
149 stable for most factors, in particular for longer-lived factors (Factor IV, VI).
150 Factor V also exhibited a quite stable behavior with only some variability of
151 ethane. The bootstrap process attributed factor profiles with various different
152 contributions of aromatic compounds, *n*-heptane, and, to some extent,
153 methylpentanes and hexane, to both Factors II and III. For Factor II the median of
154 the contributing percentages of *n*-heptane, toluene, ethylbenzene, and xylenes of
155 the 200 bootstrap profiles was much higher than the base run value. The stabilities
156 from the bootstrap runs with the inclusion of the photochemical aging showed a
157 tendency that factors with high relative contributions of long-lived substances
158 were more stable than those with high contributions of substances with higher rate
159 constants such as ethene, propene, and toluene, as expected.

160 In addition to the less stable bootstrap runs, transport times from major source
161 areas such as the surrounding cities of Munich, Augsburg, and Weilheim were on
162 average shorter than twelve hours. Additionally, the patterns of the factors were
163 shifted towards longer-lived compounds, reducing the influence of species with
164 higher reactivity in a way that the new factors do not well resemble the original
165 emission profiles at the sources. Aging can thus better be taken into account when
166 the profiles are interpreted rather as aged profiles than source emission profiles
167 unaltered by higher uncertainties due to photochemical reactivity. In conclusion,
168 an application of the method proposed by Sauvage et al. (2009) seems not suitable
169 to better interpret the results at this site.

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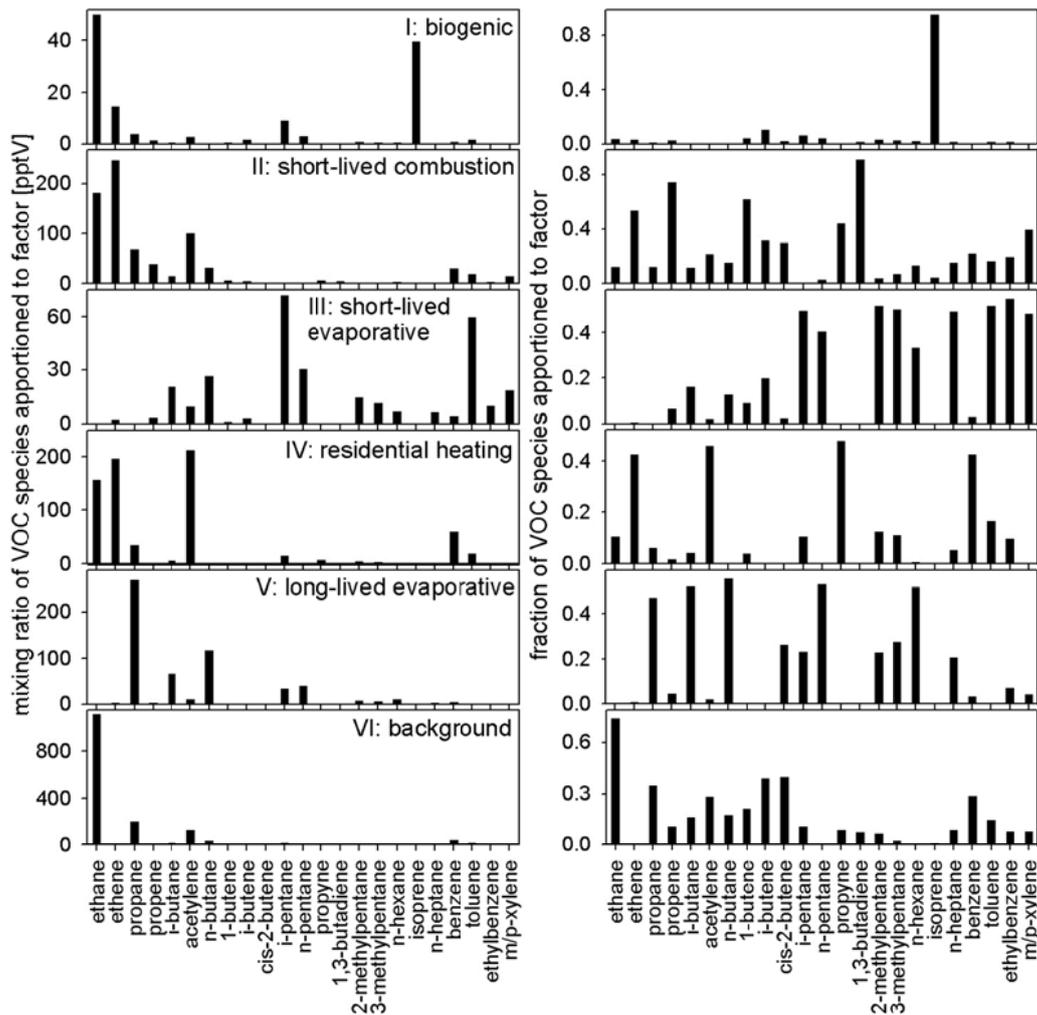
172 Table S1. Mean values of calculated uncertainties for assumed time of reaction

173 with OH of 0 hours and 36 hours.

Compound	Average uncertainty [%]	
	0 h	36 h
Ethane	7.1	8.0
Ethene	11.0	56.3
Propane	6.2	13.5
Propene	12.6	82.7
Isobutane	6.9	21.6
Acetylene	21.0	25.2
<i>n</i> -butane	26.0	39.2
1-butene	29.3	92.4
Isobutene	16.9	96.2
<i>cis</i> -2-butene	39.1	105.3
Isopentane	11.2	33.8
<i>n</i> -pentane	6.8	32.4
Propyne	33.5	55.1
1,3-butadiene	43.6	108.9
2-methylpentane	15.0	42.8
3-methylpentane	25.7	48.9
<i>n</i> -hexane	18.3	45.3
Isoprene	33.3	115.8
<i>n</i> -heptane	12.9	48.4
Benzene	8.6	16.3
Toluene	10.9	44.8
Ethylbenzene	30.7	59.5
<i>m+p</i> -xylene	25.2	72.8

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Figure S5. Factor profiles for the six factor solution calculated by PMF, assumed reaction time: 36 h, for isoprene: 0 h. left: mixing ratio of each species apporportioned to each factor [pptv], right: contribution of each factor to the species.

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S6. Photochemical age-based parameterization method

S6.1 Methods

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The parameterization method uses the ratio of two NMHC that react at different rates with OH to represent photochemical age of an air mass and thus photochemical reaction time (de Gouw et al., 2005). OH exposure as product of the OH radical concentration [OH] and reaction time Δt was determined from Eq. (S1).

189

$$[OH]\Delta t = \frac{1}{(k_A - k_B)} \times \left[\ln \frac{[A]}{[B]} \Big|_{t=0} - \ln \frac{[A]}{[B]} \right] \quad (\text{S1})$$

190

The parameters k_A and k_B are rate constants of the respective NMHC (Atkinson

190 and Arey, 2003), $\frac{[A]}{[B]}|_{t=0}$ is the initial ratio of the NMHC mixing ratios in fresh
 191 emissions, and $\frac{[A]}{[B]}$ the measured NMHC ratio. For this approach the ratios of
 192 benzene and toluene, *m+p*-xylene and benzene, as well as *m+p*-xylene and
 193 ethylbenzene were used for the determination of photochemical age and OH
 194 exposure (de Gouw et al., 2005; Yuan et al., 2012; 2013). The objective was to
 195 evaluate if this method is capable to help interpreting the resolved factors in terms
 196 of photochemical degradation and source apportionment rather than to determine
 197 emission ratios. However, emission ratios were calculated to check for plausibility
 198 of the method following de Gouw et al. (2005) and Yuan et al. (2012) using Eq.
 199 (S2):

$$200 \frac{[NMHC]}{[C_2H_2]} = ER \times e^{-(k_{NMHC}-k_{C_2H_2})[OH]\Delta t} \quad (S2)$$

201 Here, [NMHC] and [C₂H₂] are mixing ratios of the particular NMHC and
 202 acetylene, respectively, k_{NMHC} and $k_{C_2H_2}$ are OH rate constants, ER is the
 203 emission ratio of the respective NMHC to acetylene. OH exposure ([OH]Δt) was
 204 calculated with Eq. (S1). ER and k_{NMHC} are unknown and were determined from
 205 regression fits as described in de Gouw et al. (2005). The results and applicability
 206 of the photochemical age-based parameterization method are discussed in sections
 207 S6.2 and S.6.3

208 Yuan et al. (2012) further suggested two methods to test if the PMF factors are
 209 derived from independent sources or rather from different stages of
 210 photochemical processing from one common source. The first test is for
 211 dependency of the factor fractions of the respective factors to the OH reaction rate
 212 constant (k_{OH}). In case that a factor extracted from PMF derived from a specific
 213 source category, no dependency should be observed. The second test takes into
 214 account the relationship of two extracted factor profiles in terms of photochemical
 215 processing. The ratio of the profiles *i* and *j* ($R_{profile(i)/profile(j)}$) should follow Eq.
 216 (S3), if two factors are photochemically derived from one another with *A* being a
 217 scaling factor accounting for normalization procedures in the PMF model (Yuan
 218 et al., 2012).

$$219 R_{profile(i)/profile(j)} = A \times e^{-k_{OH}[OH]\Delta t} \quad (S3)$$

220 The results of the factor dependency tests are shown and discussed in section
 221 S6.3.

222

223 **S6.2 Results and Discussion**

224 Yuan et al. (2012) used the ratio *m+p*-xylene/benzene (M/B), Yuan et al. (2013)
225 *m+p*-xylene/ethylbenzene (M/E), and de Gouw et al. (2005) benzene/toluene
226 (B/T) for the determination of the photochemical age. In this study all three ratios
227 were applied and OH exposure and photochemical age compared to test for
228 plausibility. The differences in the resulting photochemical age were very large up
229 to a factor of 6.5 (median of photochemical age calculated for all samples) and
230 14.1 (arithmetic mean) between the three methods with large outliers at certain
231 times of factors up to 1000. A reliable calculation of photochemical age out of the
232 ratio of two compounds with different reactivity was not possible with our dataset
233 and thus the derived emission ratios also varied largely depending on the NMHC
234 ratio applied for the photochemical age calculation. The choice of a single ratio
235 for the determination of the photochemical age seems quite arbitrary. This is due
236 to several properties at the site Hohenpeissenberg: we cannot hold the assumption
237 (assumption 4 from Yuan et al. (2012)) that the initial emission ratios of the
238 different contributing sources are constant for all sources or air masses that
239 reached the site that in addition have very different reaction times with OH, since
240 the site is exposed to the impact of many overlying sources in different distances.
241 Thus, it is not possible to separate aging, mixing, and source type from one
242 specific ratio. Therefore, the photochemical age-based parameterization method is
243 not applicable at our site.

244

245 **S6.3 Evaluation of photochemical dependency of the factors**

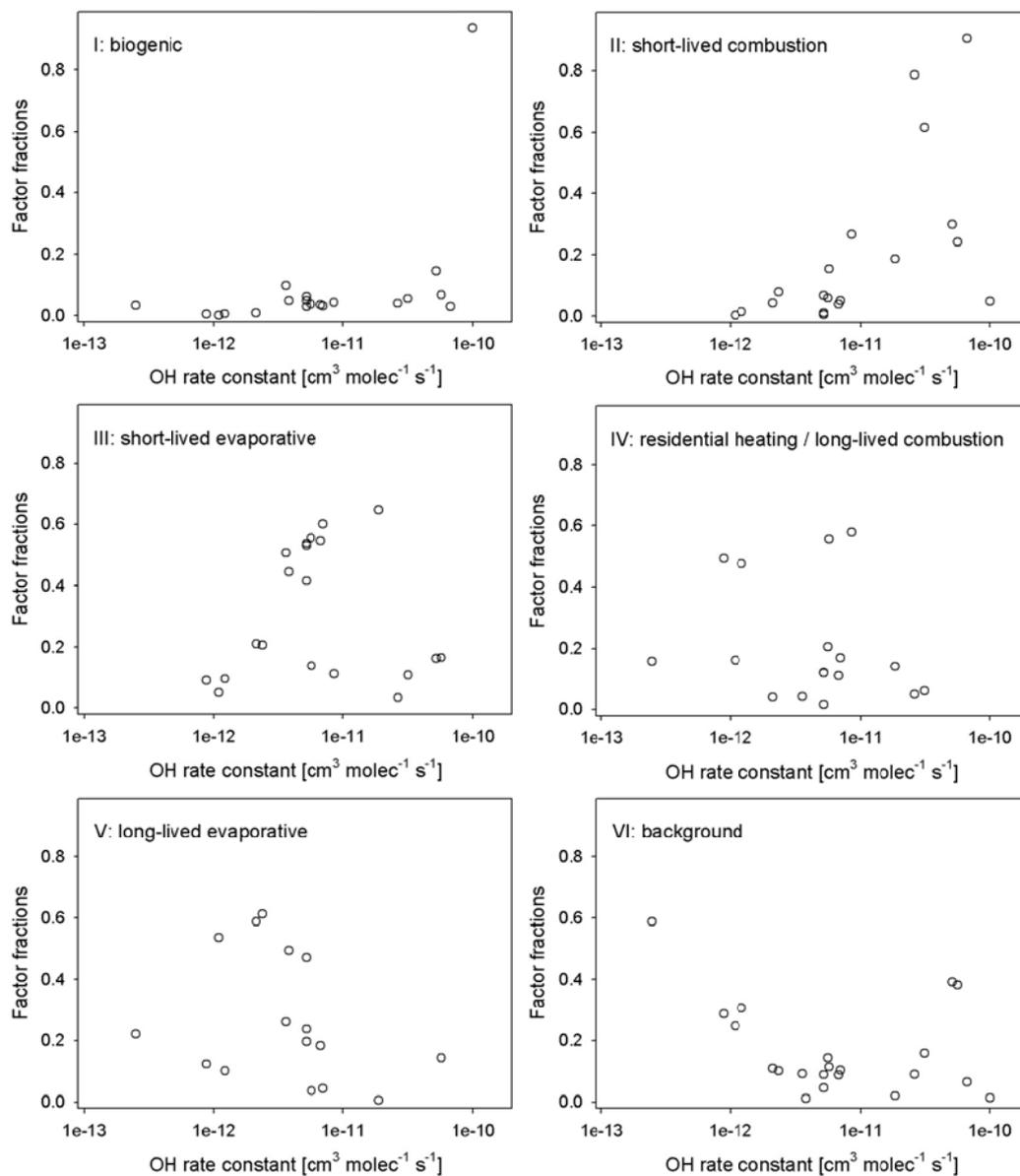
246 Yuan et al. (2012) suggested tests to check for photochemical dependency of the
247 resolved factors. In case that the factors extracted from PMF are derived mainly
248 from photochemical processing a dependency on the rate constants should be
249 seen. Figure S6 illustrates relative factor contributions to each NMHC species as a
250 function of the OH rate constants. Within the two evaporative categories (short-
251 lived and long-lived) and also the two combustion categories (short-lived and
252 residential heating/long-lived) a split of short and long-lived compounds could be
253 observed between the two respective factors. Not a clear photochemical
254 relationship, as shown by Yuan et al. (2012) for Beijing data, could be seen in the
255 Hohenpeissenberg analyses. The factors discussed here showed a certain
256 separation by photochemical processing into short and long-lived factors, but

257 were still separated into the source categories evaporative and combustion,
258 supporting the source apportionment from section 3.2.

259 The second method tests the relationship of two factor profiles in terms of
260 photochemical aging. If one factor was photochemically derived from another
261 one, a dependency of the factor ratio to the rate constants should be found. Figure
262 S7 illustrates results of the following factor ratios: background and short-lived
263 combustion, background and residential heating/long-lived combustion,
264 background and long-lived evaporative, background and short-lived evaporative,
265 long-lived evaporative and short-lived evaporative, residential heating/long-lived
266 combustion and short-lived combustion. None of the tested factor ratios was
267 explained by a relation according to Eq. (S3), not even between short-lived and
268 long-lived profiles within the evaporative or combustion categories. This indicates
269 independency of the factor profiles in term of photochemical processing.

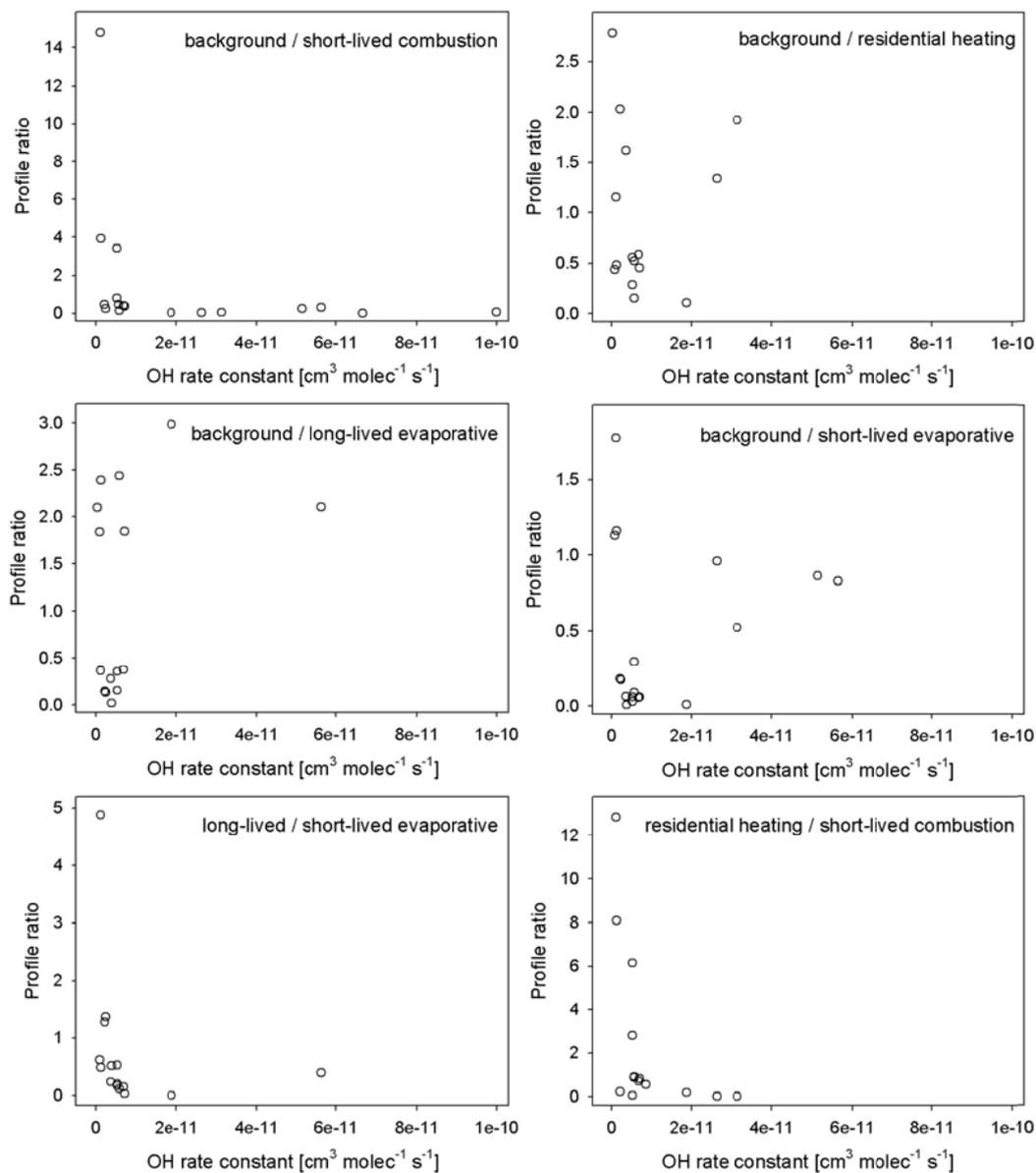
270 Both tests did not show clear signs of photochemical relationship between the
271 different extracted profiles. As has been discussed before, this is attributed to the
272 complex situation of multiple sources with different transport times overlaying in
273 the receptor profiles.

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Figure S6. Factor contributions to each NMHC species as a function of the rate constants (k_{OH}). Each data point represents one compound. No clear dependence on k_{OH} can be seen.



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Figure S7. Abundance ratios of the compounds from two profiles as a function of k_{OH} values. No dependence on k_{OH} can be seen which means that the factors are not directly photochemically derived from each other.