



Supplement of

Cloud droplet activity changes of soot aerosol upon smog chamber ageing

C. Wittbom et al.

Correspondence to: C. Wittbom (cerina.wittbom@nuclear.lu.se)

1 Supplement



2

3

Figure S 1. Illustration of calibration measurements performed with AS (for $\Delta T=4$ K, 10 K and 18 K) and sucrose ($\Delta T=18$ K) for one of the CCNC. Measured values (markers) and the corresponding fit (lines) with 95% confidence intervals (dotted lines) for the three different ΔT used in the study. The slope and intercept of the 18 K line are 0.00221984 and -0.1882434; the slope and intercept for the 10 K are 0.00146469 and -0.2063209; and for the 4 K the slope and intercept are 0.00067888 and -0.1128603.

Table S 1. An example of the DMA-APM settings (without the thermodenuder) used during 1

2 DEP2. Shown here is the particle mobility diameter (d_m) , the peak of the mass distribution for the fresh ($V_{\text{fresh soot}}$) and processed soot ($V_{\text{processed soot}}$), the rotational speed and voltage scan-3 4

range of the APM as well as the resolution parameter for the fresh ($\lambda_{\text{fresh soot}}$) and the

5	processed	soot	$(\lambda_{\text{processed soot}}).$	
---	-----------	------	--------------------------------------	--

$d_{\rm m}$	$V_{\text{fresh soot}}$	Vprocessed soot	APM Rotational	APM Voltage	$\lambda_{\text{fresh soot}}$	$\lambda_{\text{processed soot}}$
(nm)	(V)	(V)	speed	scan-range		
			(rpm)	(V)		
90	50	100	5500	20-160	0.062	0.125
150	140	420	5000	40-400	0.074	0.223
200	180	620	3000	50-500	0.061	0.210
300	260	900	3000	500-2000	0.049	0.170



Figure S 2. Empirically derived volume equivalent diameters (d_{ve}) for mobility diameters (d_m) of 90 nm (red), 150 nm (green), 200 nm (purple) and 300 nm (blue). Diesel exhausts particles (DEP1, 2 and 3, triangles) as well as flame soot particles (FSP1, circles) with a primary particle diameter (d_{pp}) of 28 nm are used for the fitting. The estimated d_{ve} (lines) is calculated from measured d_m and the SOA mass fraction (mf_{SOA}) of the particles (see Eq. (11) and (12) in Sect.5.3).

9

1 ADCHAM model simulations of the DEP2 experiment

2 Here we describe how the Aerosol Dynamics, gas- and particle- phase chemistry model for 3 laboratory CHAMber studies (ADCHAM; Roldin et al., 2014) was used to simulate the gas-4 phase chemistry, secondary organic aerosol (SOA) formation and its properties during the DEP2 experiment. For the model simulations it is assumed that the gas-phase chemistry in the 5 6 chamber is primarily driven by the NO_x emissions from the Euro II Diesel Passenger Vehicle, 7 the added light-aromatic precursors (toluene and m-xylene), and the alkene, alkyne and 8 aldehyde emissions from the diesel vehicle. The absolute alkene, alkyne and aldehyde 9 concentrations in the chamber (c_i) (Table S2), are estimated by scaling their concentrations reported by Schauer et al. (1999) ($c_{i,Schauer}$) with our measurements of light-aromatic 10 compounds (C6-C9) from the DEP4 experiment (Eq. (S1)). 11

12

13
$$C_i = C_{i,Schauer} \cdot \frac{[\text{light-arom.}]_{\text{DEP4}}}{[\text{light-arom.}]_{\text{Schauer}}}$$
 (S 1)

15 Table S 2. Estimated initial alkene, alkyne and aldehyde concentrations in the Teflon chamber

10 during the DEP2 and DEP4 experiment	16	during th	e DEP2	and DEP4	experiment
--	----	-----------	--------	----------	------------

	c_i (ppbv)	
Alkenes		
Ethene	13.03	
Propene	0.79	
Trans-2-butene	0.20	
Cis-2-butene	0.40	
Isobutene	0.87	
3-methyl-1-butene	0.10	
2-methyl-1-butene	0.16	
1,3-butadiene	0.24	
Total	15.79	
Alkynes		
Ethyne	7.54	

Aldehydes	
Formaldehyde	31.68
Acetaldehyde	40.48
Propanal	10.29
Total	82.45

2 Gas-phase chemistry

3 In the model the simulations start with estimated concentrations of alkene, alkyne and 4 aldehyde (Table S2) and the measured initial NO and NO₂ concentration of 420 and 150 ppb, 5 respectively. Between 20 and 90 minutes after the input of diesel vehicle emissions O₃ are 6 slowly added (~5.3 ppb/min) to the modelled chamber in order to capture the conversion of 7 NO to NO₂ prior to the onset of the UV-light (see Fig. S3a). 30 minutes before the UV-lights 8 are turned on 430 ppb of toluene and 310 ppb of *m*-xylene are added to the chamber. 9 Unfortunately the toluene and *m*-xylene concentrations were not measured during the DEP2 10 experiment. Thus, the concentrations were estimated based on the GC-MS measurements 11 during the DEP4 experiment, in which the same amount of toluene and *m*-xylene were 12 injected. The gas-phase chemistry in the chamber was modelled with the kinetic mechanism 13 from Master Chemical Mechanism v3.2 (MCMv3.2; Jenkin et al., 2003; Bloss et al., 2005 and 14 b), including all compounds in Table S 2, toluene, *m*-xylene and all inorganic reactions (in 15 total 772 compounds and 2446 reactions). Bloss et al. (2005a and b) have shown that the 16 MCMv3.1 generally overestimates the ozone concentration and underestimates the OH 17 concentration during oxidation of light aromatic compounds (e.g. xylene and toluene). Similar model and measurement discrepancy was also observed by Roldin et al. (2014) when 18 simulating the photooxidation of *m*-xylene in the 6 m^3 Teflon chamber used in the present 19 20 study. In order to improve the model performance Bloss et al. (2005) and Roldin et al. (2014) 21 included an artificial OH source after the UV-lights are turned on. In this work the same artificial OH source rate are used as in Roldin et al. (2014) of 10^8 cm⁻³ s⁻¹. With the artificial 22 23 OH source the model better captures the observed temporal evolution of the NO and O₃ 24 concentrations.

1 **Chamber wall effects**

2 The ADCHAM model explicitly treats the deposition of particles and gases onto the Teflon 3 walls, the mass transfer limited diffusion of gases across a thin laminar layer (Δx) adjacent to 4 the Teflon walls, and the mass transfer of organic compounds to and from the particles deposited on the chamber walls. Table S3 list the values used for different model parameters 5 6 related to the mass transfer to and from the Teflon walls. See Roldin et al. (2014) for detailed 7 information about how all these parameters are used in ADCHAM. The parameter values 8 were adopted from Roldin et al. (2014) in which ADCHAM was used to simulate the SOA 9 formation during a *m*-xylene experiment in the Teflon chamber in the Aerosol Laboratory at 10 Lund University.

Parameter	Description	Value (unit)
V_0	Initial chamber volume	5.95 (m ³)
$\Delta V / \Delta t$	Volume loss rate in chamber	$-0.4^{\#} (m^3 h^{-1})$
V_t	Chamber volume at time t	$V_0 + \Delta V / \Delta t \cdot t$
\overline{E}	Mean electrical field strength	$50 \cdot V_0 / V_t ~({\rm V~cm}^{-1})$
u*	Friction velocity	$0.25^{\#}, 0.05^{\#}** \text{ (m s}^{-1})$
$k_{g,w}$	First order loss rate from the near	$1/15 (s^{-1})$
	wall gas phase to the walls	
$k_{w,g}$	Desorption rate from the chamber	$k_{g,w}$ (s ⁻¹)
	wall Teflon surfaces	$\kappa_{w,g,i} = \frac{1}{\left(RT / p_{0,i}C_w / \left(M_w \gamma_{w,i}\right)\right)} \left(3\right)$
$C_{w,i}$	Effective wall equivalent mass	$(\text{mol } \text{m}^{-3})$
	concentration of compound <i>i</i>	
$M_{_W}$	Average molar mass of the Teflon	(mol)
	film	
$\gamma_{w,i}$	Activity coefficient of compound <i>i</i>	
	in the Teflon film	
$p_{0,i}$	Pure liquid saturation vapour	(Pa)
	pressure of compound <i>i</i>	
$C_{w}/(M_{w}\gamma_{wi})$	Measureable parameter, for uptake	100 (µmol m ⁻³)***
w (w w,t)	on Teflon walls (see Matsunaga	
	and Ziemann, 2010)	
Δx	Laminar layer adjacent to the	10^{-3} (m)
	Teflon walls	
E_{HONO}	HONO wall emissions during	4.6×10^8 molecules cm ⁻² s ⁻¹
	photooxidation	

11 Table S 3. Parameters used to calculate the chamber wall effects in ADCHAM.

12 13 [#]Different value than used by Roldin et al. (2014).

**Value used before the UV-light are turned on.

14 ***The same value was used for all condensable organic compounds



Figure S 3. Modelled (a) NO(g), (b) O₃(g), (c) OH(g) and (d) *m*-xylene(g) and toluene(g)
concentrations for the DEP2 experiment. In (a) and (b) the modelled concentrations can be
compared with the measured. The UV onset is at time 0 h in the figures.

1 Particle-phase chemistry and SOA formation

2 Oxidation products from the added light-aromatic precursors m-xylene and toluene dominate 3 the SOA formed during the DEP2 experiment. However a small fraction of the SOA is also 4 originating from other organic compounds present in the diesel exhausts (e.g. *n*-alkanes and 5 PAHs).

6 The SOA formation from *m*-xylene and toluene are modelled by considering the non-7 equilibrium gas-particle partitioning of all their non-radical oxidation products in MCMv3.2. 8 In Roldin et al. (2014) it was shown that ADCHAM could not capture the observed early 9 stage SOA formation during an *m*-xylene oxidation experiment, if not considering particle-10 phase oligomerization. However, with peroxyhemiacetal (aldehydes + hydroperoxides) and hemiacetal (alcoholes + aldehydes) dimer formation rates (k_i) of 10⁻²² cm³ s⁻¹, ADCHAM 11 were able to nearly reproduce the temporal evolution of the SOA concentration during the 12 13 complete oxidation experiment (~4 hours). Here we will test to simulate the SOA formation 14 both with and without these particle-phase oligomerization processes. 15 In order to account for the SOA formation from known PAHs and *n*-alkanes measured in diesel exhausts, we use the measurements from Schauer et al. (1999). Eq. (S1) is used for 16 17 estimating the absolute *n*-alkane and PAH concentrations in the chamber. Table S4 gives the

- 18 estimated initial PAH and *n*-alkane concentrations, and their first order reaction rate with OH
- 19 radicals $(k_{OH,i})$.
- 20

21	Table S 4. Estimated initial PAH and <i>n</i> -alkane concentrations in the Teflon chamber during
22	the DEP2 and DEP4 experiments, and $k_{OH,i}$ used in the model simulations.

	c_i (ppbv)	$k_{OH,i}$ (cm ³ molec ⁻¹ s ⁻¹)
PAHs		
Naphthalene	0.2054	2.3x10 ^{-11 (a)}
2-methylnaphthalene	0.1860	4.86×10^{-11} (b)
1-methylnaphthalene	0.1151	4.09x10 ^{-11 (b)}
C2-naphthalene	0.1520	$6.0 \times 10^{-11} ^{(c)}$
C3-naphthalene	0.0624	8.0x10 ^{-11 (d)}
C4-naphthalene	0.0236	8.0x10 ^{-11 (d)}
other PAHs	0.2010	$8.0 \times 10^{-11} $ (d)
Total	0.9453	
n-alkanes		
Dodecane	0.1276	$1.32 \times 10^{-11} $ (a)
Tridecane	0.1117	$1.51 \times 10^{-11} $ (a)
Tetradecane	0.1368	$1.79 \times 10^{-11} ^{(a)}$

Pentadecane	0.0808	2.07x10 ^{-11 (a)}
Hexadecane	0.1353	2.32×10^{-11} (a)
Heptadecane	0.1099	2.85x10 ^{-11 (e)}
Octadecane	0.1016	$3.51 \times 10^{-11} (e)$
Nonadecane	0.0658	$4.32 \times 10^{-11} ^{(e)}$
Eicosane	0.0414	$5.31 \times 10^{-11} (e)$
Total	0.9108	

^(a)Atkinson and Arey (2003)

^(b)Phousongphouang and Arey (2002)

^(c)Average of all dimethylnaphthalanes in Phousongphouang and Arey (2002)

1 2 3 4 5 ^(d)Estimated from dimethylnaphthalenes, assumed to be same for all other PAHs according to Chan et al. (2009)

^(e)Estimated from structure-reactivity relationships (Kwok and Atkinson, 1995)

6

7 The SOA formation from the PAHs were modelled with the one or two product SOA yield 8 parameterizations from Chan et al. (2009), for high-NO_x conditions. For the n-alkanes we 9 estimate the SOA yields based on the modelling from Jordan et al. (2008), using a 1-product model parameterization (Eq. (S2)), 10

11
$$Y_i = M_o \frac{\alpha_i K_i}{1 + K_i M_o}$$
(S 2)

12 where α_i is the mass-based stoichiometric yield of the oxidation product *i*, K_i is its gas-particle

13 partitioning equilibrium constant, and M_o is the total particulate organic mass concentration

14 (listed in Table S5).

15

16 Table S 5. Estimated mass-based stoichiometric yields (α_i), equilibrium partitioning constants

17 (K_i) and pure liquid saturation vapour pressures $(p_{0,i})$, derived from the SOA yield

18 parameterizations of different n-alkanes (Eq. (S2)).

c_i	α_i	$K_i (m^3 \mu g^{-1})$	$p_{0,i}$ (Pa)
Dodecane	0.0472	0.0489	2.109×10^{-4}
Tridecane	0.0898	0.0333	2.925×10^{-4}
Tetradecane	0.1536	0.0320	2.883×10^{-4}
Pentadecane	0.2775	0.0272	3.221×10^{-4}
Hexadecane	0.3531	0.0341	2.447×10^{-4}
Heptadecane	0.4402	0.0554	1.437×10^{-4}
Octadecane	0.5430	0.0823	9.25x10 ⁻⁵
Nonadecane	0.6280	0.1350	5.41×10^{-5}
Eicosane	0.7837	0.1668	4.20×10^{-5}

ADCHAM includes a kinetic multilayer model, which considers the diffusion of compounds 19

20 between different particle layers. For the model simulations performed here each particle is

21 treated as composed of one solid soot core, one amorphous organic bulk phase layer and one

22 particle surface monolayer, with limited diffusion between the layers. The aerosol dynamic

- processes considered are: Brownian coagulation, condensation/evaporation and dry deposition
 to the chamber walls. For more details see Roldin et al. (2014).
- 3 Before the onset of UV radiation in the chamber a substantial increase of the O:C ratio from ~0.05 to ~0.2 is observed, meanwhile the H:C ratio decreases from ~1.9 to ~1.75 (see Fig. 4 5 S4d). However, no particle mass increase or change in the mass spectra is observed during this time, other than a slight increase in m/z 44 due to CO_2^+ . According to the ADCHAM 6 model simulations there were very low concentrations of NO₃ radicals ($<10^6$ molecules cm⁻³), 7 OH (\leq 3000 molecules cm⁻³) and O₃ (\leq 1 ppbv), before the UV-light was turned on. Hence, the 8 9 change in O:C and H:C ratio during the dark conditions is unlikely attributed to SOA 10 formation. Instead it can be explained by heterogeneous oxidation of POA with NO₂, which reaches a maximum concentration of ~500 ppb, just before the onset of the UV-lights. The 11 12 heterogeneous oxidation of diesel soot coated with POA has previously primarily been 13 studied because of the potential importance for HONO formation (e.g. Arens et al., 2001; Han 14 et al., 2013).
- 15 For the simulations presented here it is assumed that the POA is composed of slightly 16 oxidized *n*-alkanes, and alkenes with an average number of carbon atoms per molecule of 26, 17 two carbon-carbon double bonds, and one ketone functional group (H:C=1.8846, O:C=0.0385). The POA are assumed to react with NO₂ with a reaction rate (k_{NO_2}) equal to 18 10⁻¹⁸ cm⁻³ s⁻¹, forming oxidized POA (OPOA) with an H:C of 1.6538 and O:C of 0.1923. The 19 20 uptake of NO₂ is modelled with the kinetic multilayer model, considering the adsorption and 21 diffusion of NO₂ between the surface monolayer and the organic bulk phase, analogous to Roldin et al. (2014). The NO₂ diffusion coefficient was assumed to be equal to 10^{-8} cm² s⁻¹. 22

23 Figure S4 shows the modelled and measured: (a) particle number concentration, (b) total 24 organic particle mass concentration in the air, (c) organic mass fraction (mf_{OA}) and (d) H:C 25 and O:C ratios. When accounting for deposition of charged and neutral particles two the 26 chamber walls, according to the procedure described in Roldin et al. (2014) (although with 27 five times larger friction velocity) and Brownian coagulation, the modelled particle number 28 concentrations are in good agreement with the observations. The model performance 29 substantially improves if we consider hemiacetal and peroxyhemiacetal (oligomer) formation. 30 However, in contrast to the *m*-xylene experiment in Roldin et al. (2014), ADCHAM still substantially underestimates the total organic particle mass and mf_{OA} (SOA formation) 0-2 31 32 hour after the UV exposure starts.

1 Figure S 5 shows the temporal evolution of the modelled mass concentration of POA, 2 OPOA and SOA formed from oxidation products of light-aromatic compounds (m-xylene and 3 toluene), *n*-alkanes and PAHs. According to the model results the mass contribution from *n*-4 alkane oxidation products are negligible, and the contribution from the considered PAHs 5 (Table S3) are ~4 times too small to explain the observed early stage SOA formation. This 6 indicates that we either substantially underestimates the PAH concentrations in the chamber 7 or that there are some unknown IVOCs in the Diesel exhausts which we do not account for. 8 Another possibility could also be that we cannot realistically represent the early stage SOA 9 formation from the *m*-xylene and toluene oxidation products. Possibly, a reactive uptake 10 mechanism not directly driven by the oxidation products saturation vapour pressures but their 11 reactivity with other organic compounds on the particle surfaces, could explain the almost 12 immediate onset of the SOA formation, after the UV-light is turned on. However, this is not 13 fully consistent with the *m*-xylene experiment simulated in Roldin et al. (2014), or other 14 classical photooxidation experiments of light-aromatic compounds (see e.g. Ng et al., 2007). 15 Figure S6 shows the measured and modelled total organic particle mass plotted against the amount of reacted *m*-xylene and toluene (Δm -xylene + Δ toluene). According to the model 16 simulations ~120 μ g/m³ of *m*-xylene and ~40 μ g/m³ of toluene have been consumed before 17 the OA mass starts to increase in the air. However from the measurements in combination 18 19 with the modelled *m*-xylene and toluene decay, only $\sim 30 \ \mu g/m^3$ of *m*-xylene and $\sim 10 \ \mu g/m^3$ of toluene need to react before the OA mass increases. 20



1 2

Figure S 4. Modelled and measured (a) particle number concentration, (b) OA mass, (c) mf_{OA} and (d) H:C and O:C ratios for the DEP2 experiment. The onset of UV exposure is at time 0 h in the figures. In (b) and (c) the model results from simulations with (solid line) and without (dashed line) particle phase oligomerization are included. In (a) and (d) the results are from the simulation with oligomerization.



1 2

Figure S 5. Modelled (lines) and measured (markers; APM-SMPS) total organic particle mass (in the air) for the DEP2 experiment. Given are also the modelled POA (turquoise), OPOA (pink) formed from the heterogeneous reactions between POA and NO₂, and SOA originating from the light-aromatic precursors (m-xylene and toluene) (red), PAHs (blue) and *n*-alkanes (black). We have also included the modelled total organic mass (light green) in the air and on the wall deposited particles. At time 0 h in the figure the UV-lights are turned on.



Figure S 6. Measured (APM) and modelled total organic particle mass concentration (*OA_{tot}*)
plotted against the amount of reacted *m*-xylene and toluene.



3 Figure S 7. Comparison of calculated s_c and empirical results (a) and comparison of κ_{SOA} with 4 κ_{CCN} (b), calculated from the chemical composition and measured s_c respectively. Colour code 5 with respect to particle mobility diameter (d_m) and different markers with respect to type of 6 soot generator (experiment FSP1, 2 or DEP2). The large discrepancies for the high values of 7 s_c in (a) are due to the under prediction of the s_c by the model of the early aged soot particles. 8 Small values in (b) correspond to modelled and measured values (κ_{SOA} and κ_{CCN}) for early 9 aged soot particles, where uncertainties are large for κ_{CCN} . For large values in (b) the uncertainties are inherited from empirical fitting of the mf_{SOA} (visible for FSP1; $d_m = 150$ and 10 11 300 nm, and for FSP2; $d_{\rm m}$ =60 and 90 nm).

13 Determining mf_{OA} from the SP-AMS data

14 Quantification by means of AMS is a two-step process. First, the signal is assigned to species 15 (such as 'organics', 'rBC' etc., see next paragraph), using the fragmentation table approach 16 (Allan et al., 2004). Then, a species is quantified using two parameters: collection efficiency 17 (CE) and ionization efficiency (IE) (Allan et al., 2004). CE refers to the fraction of the mass 18 that is eligible for ionization, normally dominated by bounce of the tungsten vaporizer 19 (Huffman et al., 2005). For PM vaporized by the SP-module, the main issue is overlap of 20 particle and laser beams. IE is the probability that ionization occurs, decoupled from CE (i.e. 21 given that the material is not bouncing of the tungsten vaporizer, missing the laser beam etc.) Since ammonium nitrate is the default calibration substance, IEs are often expressed in terms
 of relative ionization efficiencies compared to nitrate.

3 Refractory Black Carbon (rBC) mass loadings were estimated assuming a relative ionization efficiency (compared to nitrate) of 0.2, applied to the C_{1-9}^+ ions (signal due to $C_{>9}$ was 4 negligible) after accounting for the organic contribution to C_{1+}^{+} (the contribution to C_{2-9+}^{+} was 5 negligible). C_1^+ was apportioned to rBC as 70% of C_3^+ , (C_3^+) was chosen because it is the most 6 7 abundant) the remaining signal from C_1^+ was interpreted as organic. This was based on measurements on fresh soot in each experiment, where the organic contribution to C_1^+ was 8 negligible. rBC also generated CO_2^+ ions (and presumably CO^+ , but these were not retrieved 9 owing to the interference from N_2^+), these were found to correspond to 70-90% of C_3^+ . These 10 CO_2^+ ions where excluded from the results presented here, pending further investigation. A 11 CE of 1 was applied for both organics and rBC. 12

13

14 Determining mf_{OA} from combined DMA-APM and SMPS measurements

The total organic mass fraction of the polydisperse particle size distribution was determined 15 16 from a procedure that started with determining the mass size distribution by multiplying the 17 measured number size distribution from the SMPS with the size resolved mass per particle 18 determined with the DMA-APM. This mass size distribution was then divided into a volatile 19 (organic) and a non-volatile fraction by, in a size resolved manner, multiplying the mass 20 distribution with the volatile mass fraction for each size bin. The DMA-TD-APM measured in 21 the size range 60 to 300 nm at five discrete mobility sizes. Data at other sizes were obtained 22 by fitting a physically relevant function.

23

24 Corresponding atmospheric ageing timescales

The modelled cumulative OH exposure during the end of the DEP2 experiment is 7.7x10⁶ cm⁻ ³ h. This together with the measured mass spectral signature of the organic material (see Sect. 5.2) illustrate that, with respect to chemical composition, the SOA during the experiment is relatively fresh (a few hours of ageing for typical summer daytime conditions). Still, for midlatitude winter conditions this cumulative OH exposure may very well correspond to one or a few days in the atmosphere.

In the smog chamber experiments the SOA precursor concentration is substantially higher than for typical atmospheric conditions, while the condensation sinks of the diesel soot particles, are comparable with typical urban plume conditions (see e.g. Roldin et al., 2011). Hence, substantially higher mass growth rates are expected in the chamber compared to
 typical atmospheric urban plume conditions. To some extent this is counteracted by the
 uptake of condensable organic compounds onto the Teflon walls and wall deposited particles.

4 In the smog chamber the time of photochemical ageing, before the soot particles become 5 CCN active at a supersaturation of 0.2%, range from 1.5 to >4.5 h depending on experimental 6 conditions as well as particle size. To estimate the corresponding atmospheric ageing time the 7 range of observed new particle growth rates (GR) in the mid-latitudes of 1-20nm/h has been 8 used (Kulmala et al., 2004). It is also assumed that these growth rates (if converted to mass 9 growth rates) are valid for soot particles with a mobility diameter (d_m) of ~100 nm (volume equivalent diameter ($d_{ve, fresh}$) of ~70 nm). According to the experiments these particles need 10 11 to have an mf_{SOA} of ~90 % before they activate at a supersaturation of 0.2% (see Figure 11). At this stage these particles have a volume equivalent diameter $(d_{ve,aged})$ equalling the 12 13 mobility diameter ($d_{\rm m}$) of ~150 nm, i.e. the particles are almost spherical. Hence, based on this the atmospheric ageing time with respect to organic condensational growth $(t_{\rm atm})$ is 14 15 estimated to be between 4 hours and 3 days (Eq. (S3)).

$$16 t_{atm} = \frac{\left(d_{ve,aged} - d_{ve,fresh}\right)}{GR} (S 3)$$

$$17 18 19 20$$

1 **References Supplementary**

2

3 Allan, J. D., Delia, A. E., Coe, H., Bower, K. N., Alfarra, M. R., Jimenez, J. L., Middlebrook, 4 A. M., Drewnick, F., Onasch, T. B., Canagaratna, M. R., Jayne, J. T. and Worsnop, D. R. "A 5 generalised method for the extraction of chemically resolved mass spectra from aerodyne aerosol mass spectrometer data." Journal of Aerosol Science 35, no. 7 (Jul 2004): 909-922. 6

7 Arens, F., Gutzwiller, L., Baltensperger, U., Gaggeler, H. W. and Ammann, M. 8 "Heterogeneous reaction of NO2 on diesel soot particles." Environmental Science & 9 Technology 35, no. 11 (Jun 1 2001): 2191-2199.

10 Atkinson, R. and Arey, J. "Atmospheric degradation of volatile organic compounds." 11 Chemical Reviews 103, no. 12 (Dec 2003): 4605-4638.

12 Bloss, C., Wagner, V., Bonzanini, A., Jenkin, M. E., Wirtz, K., Martin-Reviejo, M. and

13 Pilling, M. J. "Evaluation of detailed aromatic mechanisms (MCMv3 and MCMv3.1) against

- 14 environmental chamber data." Atmospheric Chemistry and Physics 5 (Mar 1 2005a): 623-639.
- 15 Bloss, C., Wagner, V., Jenkin, M. E., Volkamer, R., Bloss, W. J., Lee, J. D., Heard, D. E.,

Wirtz, K., Martin-Reviejo, M., Rea, G., Wenger, J. C. and Pilling, M. J. "Development of a 16

17 detailed chemical mechanism (MCMv3.1) for the atmospheric oxidation of aromatic

18 hydrocarbons." Atmospheric Chemistry and Physics 5 (Mar 1 2005b): 641-664.

19 Chan, A. W. H., Kautzman, K. E., Chhabra, P. S., Surratt, J. D., Chan, M. N., Crounse, J. D., 20 Kurten, A., Wennberg, P. O., Flagan, R. C. and Seinfeld, J. H. "Secondary organic aerosol 21 formation from photooxidation of naphthalene and alkylnaphthalenes: implications for 22 oxidation of intermediate volatility organic compounds (IVOCs)." Atmospheric Chemistry 23 and Physics 9, no. 9 (2009): 3049-3060.

24 Han, C., Liu, Y. C. and He, H. "Role of Organic Carbon in Heterogeneous Reaction of NO2 25 with Soot." Environmental Science & Technology 47, no. 7 (Apr 2 2013): 3174-3181.

26 Huffman, J. A., Jayne, J. T., Drewnick, F., Aiken, A. C., Onasch, T., Worsnop, D. R. and 27 Jimenez, J. L. "Design, modeling, optimization, and experimental tests of a particle beam width probe for the aerodyne aerosol mass spectrometer." Aerosol Science and Technology 28 39, no. 12 (Dec 2005): 1143-1163. 29

30 Jenkin, M. E., Saunders, S. M., Wagner, V. and Pilling, M. J. "Protocol for the development 31 of the Master Chemical Mechanism, MCM v3 (Part B): tropospheric degradation of aromatic 32 volatile organic compounds." Atmospheric Chemistry and Physics 3 (Feb 12 2003): 181-193.

33 Jordan, C. E., Ziemann, P. J., Griffin, R. J., Lim, Y. B., Atkinson, R. and Arey, J. "Modeling SOA formation from OH reactions with C-8-C-17 n-alkanes." Atmospheric Environment 42, 34 35 no. 34 (Nov 2008): 8015-8026.

- 36 Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K.
- 37 R., Altieri, K. E., Mazzoleni, L. R., Wozniak, A. S., Bluhm, H., Mysak, E. R., Smith, J. D., Kolb, C. E. and Worsnop, D. R. "Carbon oxidation state as a metric for describing the
- 38
- 39 chemistry of atmospheric organic aerosol." Nature Chemistry 3, no. 2 (Feb 2011): 133-139.

- 1 Kulmala, M., Vehkamaki, H., Petaja, T., Dal Maso, M., Lauri, A., Kerminen, V. M., Birmili,
- 2 W. and McMurry, P. H. "Formation and growth rates of ultrafine atmospheric particles: a
- 3 review of observations." Journal of Aerosol Science 35, no. 2 (Feb 2004): 143-176.
- 4 Kwok, E. S. C. and Atkinson, R. "Estimation of Hydroxyl Radical Reaction-Rate Constants
- 5 for Gas-Phase Organic-Compounds Using a Structure-Reactivity Relationship an Update.".
- 6 Atmospheric Environment 29, no. 14 (Jul 1995): 1685-1695.
- 7 Matsunaga, A. and Ziemann, P. J. "Gas-Wall Partitioning of Organic Compounds in a Teflon
- 8 Film Chamber and Potential Effects on Reaction Product and Aerosol Yield Measurements."
- 9 [In English]. Aerosol Science and Technology 44, no. 10 (2010): 881-892.
- Ng, N. L., Kroll, J. H., Chan, A. W. H., Chhabra, P. S., Flagan, R. C. and Seinfeld, J. H.
 "Secondary organic aerosol formation from m-xylene, toluene, and benzene." Atmospheric
 Chemistry and Physics 7, no. 14 (2007): 3909-3922.
- 12 Chemistry and Physics 7, no. 14 (2007): 3909-3922.
- Phousongphouang, P. T. and Arey, J. "Rate constants for the gas-phase reactions of a series of
 alkylnaphthalenes with the OH radical." Environmental Science & Technology 36, no. 9
 (May 1 2002): 1947-1952.
- 16 Rissler, J., Messing, M. E., Malik, A. I., Nilsson, P. T., Nordin, E. Z., Bohgard, M., Sanati, M.
- 17 and Pagels, J. H. "Effective Density Characterization of Soot Agglomerates from Various
- 18 Sources and Comparison to Aggregation Theory." Aerosol Science and Technology 47, no. 7
- 19 (Jul 1 2013): 792-805.
- Roldin, P., Eriksson, A. C., Nordin, E. Z., Hermansson, E., Mogensen, D., Rusanen, A., Boy,
 M., Swietlicki, E., Svenningsson, B., Zelenyuk, A., and Pagels, J. "Modelling nonequilibrium secondary organic aerosol formation and evaporation with the aerosol dynamics,
 gas- and particle-phase chemistry kinetic multilayer model ADCHAM", Atmos. Chem. Phys.,
 14, 7953–7993, doi:10.5194/acp-14-7953-2014, 2014.
- 25
- 26 Roldin, P., Swietlicki, E., Massling, A., Kristensson, A., Londahl, J. L., Eriksson, A., Pagels,
- 27 J. and Gustafsson, S. "Aerosol ageing in an urban plume implication for climate."
- Atmospheric Chemistry and Physics 11, no. 12 (2011): 5897-5915.
- 29 Schauer, J. J., Kleeman, M. J., Cass, G. R. and Simoneit, B. R. T. "Measurement of emissions
- 30 from air pollution sources. 2. C-1 through C-30 organic compounds from medium duty diesel
- 31 trucks." Environmental Science & Technology 33, no. 10 (May 15 1999): 1578-1587.