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Summertime photochemistry during CAREBeijing-2007: RO_x budgets and O₃ formation

- 5 Z. Liu^{1*} Y. Wang¹ D. Gu¹ C. Zhao^{1,**} L. G. Huey¹ R. Stickel¹ J. Liao¹ M. Shao² T. Zhu² L. Zeng²
 6 A. Amoroso³ F. Costabile⁴ C.-C. Chang⁵ and S.-C. Liu⁵
- 7 [1]{School of Earth and Atmospheric Science, Georgia Institute of Technology, Atlanta, GA, USA}
- 8 [2]{College of Environmental Sciences and Engineering, Peking University, Beijing, China}
- 9 [3] {Institute for Atmospheric Pollution, National Research Council (CNR-IIA), Rome, Italy}
- 10 [4] {Institute for Atmospheric Sciences and Climate (ISAC), CNR, Rome, Italy}
- 11 [5]{Research Center for Environmental Changes (RCEC), Academic Sinica, Taipei, China}
- 12
- 13 [*] {Now at Combustion Research Facility, Sandia National Laboratories, Livermore, CA, USA}
- 14 [**] {Now at the Pacific Northwest National Laboratory, Richland, Washington, USA}
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- 16 Correspondence to: Z. Liu (zhen.liu@eas.gatech.edu)
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18 Descriptions of instruments and experimental methods

19 O₃ and CO measurements. O₃ and CO were measured with commercial instruments 20 (ECOTECH, EC9810 and EC9830). The EC9810 Ozone Analyzer combines microprocessor control 21 with ultraviolet (UV) photometry to measure O₃ with a detection limit of 0.5 ppbv. The EC9830 22 series of Carbon Monoxide analyzers use NDIR gas filter correlation photometry to measure 23 Carbon Monoxide (CO). Zero signals were routinely measured every 2 hours by supplying dry 24 purified air into the sample line. We note that as water vapor interferes the CO analyzer, using dry 25 air will give lower baseline values, thus higher CO concentration. The uncertainties for these 26 measurements are estimated to be 5% (1 standard deviation, and so for all other uncertainties).

27 NO and NO_v measurements. NO mixing ratios were measured with a custom 28 chemiluminescence detector (SI). The instrument was calibrated periodically (2–4 hours) by a 29 standard addition of a known amount of nitric oxide. Background levels were obtained periodically 30 (30 minutes) by switching the sampled flow through a pre-reactor. NO_v levels were measured by 31 use of a molybdenum converter operated at 300. The sample flow to the NO detector was 32 periodically switched between ambient and the converter by a three way Teflon valve. NO_v 33 conversion levels were estimated to be greater than 97% for NO₂. The conversion efficiency for 34 NO_2 were examined before and after the campaign and appears to be quantitative i.e. > 95%. The 35 conversion efficiency for HNO₃ was found to be greater than 85% after the mission. We did not 36 look at anything in Beijing during the measurements, as we didn't have the ability to do so. High 37 conversion efficiency for PAN (greater than 95%) can be expected because PAN will decompose to 38 NO₂ readily at high temperature. The efficiency for NO should be comparable with NO₂, too. 39 Organic nitrates were probably close to 100% converted. The uncertainties for NO and NO_v 40 measurements are estimated to be 5% and 10%, respectively.

41 PAN measurements. Peroxyacetyl nitrate (PAN) were measured using a chemical 42 ionization mass spectrometer (CIMS) (S2, S3). The instrument was calibrated periodically with a 43 photolytic PAN source and the background response of the instrument was determined by 44 periodically scrubbing sampled air through hot stainless steel tubing (200 °C) (S4). The sensitivity 45 of the instrument was found to decrease at high NO_x levels. High levels of NO can lead to a 46 decrease in sensitivity to the CIMS PAN measurements. For the conditions in Beijing, 10 ppbv of 47 NO decreased the PAN sensitivity by 10% with higher levels decreasing the sensitivity further. NO 48 levels above 10 ppbv were observed 20% of the time but only at night. Consequently, there is 49 essentially no correction for the impact of NO during the day on PAN sensitivity. The measurement 50 uncertainty of PAN at night during high levels of variable NO is larger than during the day. 51 Detection limit for PAN is 7 pptv for 1 second integration period and a signal to noise ratio of 3. 52 The uncertainties for PAN measurements are estimated to be 10%.

HONO measurements. HONO was measured with a liquid coil scrubbing/UV-vis instrument (*S5, S6*). Briefly, gaseous HONO was trapped quantitatively in a 10-turn coil sampler using 1mm phosphate buffer. The scrubbing solution was then derivatized with sulfanilamine (SA)/N-(1-naphtyl)-ethylendiamine (NED), subsequently analyzed using high-performance liquid chromatography (HPLC), and detected by UV-vis absorption.

Interferences during sampling were studied in our laboratory in Rome (*S6*) and directly at Beijing using sodium carbonate denuders (*S7*). We used the denuder respectively in front of the Teflon tube and before the trap coils. The denuder in front of Teflon tube was used to study the possible formation of HONO from NO₂ and humidity on the tubing walls (*S8*), while the denuder before the inlet traps was used to study the nitrite formation in the water solution due to the reaction of NO₃⁻⁺SO₃ \rightarrow NO₂⁻⁺SO₄²⁻. Under summer conditions at Beijing we observed interferences on the order of 2–9% of the observed HONO mixing ratio. The inlet tube was 70 cm and the residence time of the air in the tube was less than 2 sec. The detection limit of the HONO instrument is less than 0.8 pptv. The uncertainty for HONO measurements (after removing known interference) is estimated to be 10%.

68 **VOC measurements.** C₃-C₉ NMHCs were continuously measured with 30 minute time 69 resolution, using a combination of two online GC-FID/PID systems (Syntech Spectra GC-FID/PID 70 GC955 series 600/800 VOC analyzer) (S9), one for the C_3 - C_5 NMHCs, using a gas chromatograph 71 with pre-concentration on Carbosieves SIII at 5 °C, followed by thermal desorption and separation 72 on a capillary film column and a capillary PLOT column, and quantification by a photo ionization 73 detector (PID) and a flame ionization detector (FID). The other system is for C_6-C_9 NMHCs 74 analysis; air samples are pre-concentrated on Tenax GR at normal temperature, thermal desorbed, 75 separated on an ATTM-1 column, and detected by a PID. For each analysis, an air sample with a 76 volume of 250mL was sampled. Calibration was performed before and after the campaign by using 77 a gas standard containing 39 target species with mixing ratios of 1ppm in nitrogen, prepared by the 78 gravimetric method (Spectra gases, Restek Corporation, USA). The detection limits are estimated to 79 be 10 to 90 ppty, and the uncertainty is estimated to be 5%.

Another Automated GC/MS/FID system constructed of pre-concentrator with Varian 3800 GC and Saturn 2200 MS was also deployed to measure VOCs. On each day two samples were collected and measured (8:00-9:00 and 13:00-14:00). To encompass VOCs of a wide range of volatility within each analysis, the system uses dual-columns and dual-detectors to simultaneously analyze both low and high-boiling compounds with each injection. The PLOT column connected to a FID was responsible for separation and detection of C2–C4 compounds, and the DB-1 column was connected to the MS for separation and detection of MTBE and C4–C10 compounds. Each 87 aliquot of 190 ml from the canisters was drawn to the cryogenic trap packed with fine glass beads 88 cooled at -170 °C for pre-concentration. During injection, the trap was resistively heated up to 80 89 °C within seconds, and a stream of high purity He flushed the trapped VOCs onto the columns. The 90 oven temperature was initially held at -50 °C for 3.1 min, then ramped to -10 °C at 20 °C/min, to 91 120 °C at 5 °C/min, to 180 °C at 20 °C/min, and held at 180 °C for 21.5 min. The precision of the 92 system was examined by repeatedly injecting a standard mixture made from Scott Marrin Company. 93 In general, the precision for the C2–C10 NMHCs were usually below 3%. Linearity was tested by 94 trapping a series of the same standard mixture of various concentrations (0.2-30ppbv). Most 95 compounds exhibited good linearity with RSD of calibration response factors for measured species bellow than 10%. 96

97 OVOC compounds were measured using a newly developed PFPH-GC/MS method (S10). 98 Three types of gaseous carbonyls, including formaldehyde, acetaldehyde and acetone were collected 99 within 3-h sampling period onto an adsorbent (Tenax TA) coated with pentafluorophenyl hydrazine 100 (PFPH) followed by thermal desorption and gas chromatographic (GC) analysis of the PFPH 101 derivatives with mass spectrometric (MS) detection (Agilent, GC/MS, 6890/5973N). All of the 102 tested carbonyls are shown to have method detection limits (MDLs) of sub-nanomoles per sampling 103 tube, corresponding to air concentrations of < 0.3 ppbv for a sampled volume of 24 L. These limits 104 are 2-12 times lower than those that can be obtained using the DNPH/HPLC method (S10). The 105 uncertainty for OVOC measurements is estimated to be 10%.

106 Additional descriptions of the 1-D REAM model

107 **Oxidation mechanism of aromatics.** We adopt the aromatics-oxidation mechanism in the 108 SAPRC-07 chemical mechanism developed for chemical transport models (*S14*). This mechanism is 109 chosen because careful comparison between different mechanisms (e.g. SAPRC-99, 07, CB4, RACM and the mechanism in REAM) indicates that SAPRC-07 mechanism is the most compatible with REAM and reflects the most recent updates of understanding of aromatics chemistry. In Table S1, we summarize the OH reactions for several key aromatic species, based on which the reactions for the lumped species (ARO1, ARO2) are derived. Among the products, only methylglyoxal (MGLY), glyoxal (GLYX), and biacetyl (BACL)) are shown because they are precursors for acetyl peroxy radical, which is of our interest in this work.

Model VOC input. Individual NMHC compounds were lumped into a number of model species to be used as model input. The mean concentrations for all the individual compounds and their corresponding lumped species are shown in Table S2. For species only measured at 8:00–9:00 and 13:00–14:00, interpolations were conducted based on ratios between these compounds and other continuously measured compounds with similar lifetimes.

121 Estimation of Model errors. The errors associated to the model are estimated as follows. 122 The uncertainty associated with chemical mechanism is estimated to be less than 10% based on a 123 sensitivity test, in which using different chemical mechanisms (RACM or REAM) only results in no 124 more than 10% difference of simulated PAN. The uncertainty associated with transport is estimated 125 to be 10% at most, because 5 times the modeled diffusion coefficient would only result in less than 126 10% change of PAN at surface. Taking into account all these aspects, the total error in the model is 127 estimated to be no more than 15% assuming transport and chemistry errors are not correlated and 128 not accounting for the measurement uncertainties of precursors.

129 Integration of production or loss rate and OPE in the planetary boundary layer (PBL)

130 The hourly average production or loss rate integrated in the PBL is computed using the131 following equation,

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$$\overline{x}^{j} = \frac{\sum_{i} x_{i}^{j} \cdot z_{i}^{j}}{h_{j}}$$
(S1)

where *i* is the layer index, *j* is hour index, x_i^j is the hourly production or loss rate in layer *i* and hour 133 *j*, z_i^j is the hourly thickness of layer *i* in hour *j*, and h_j is the hourly PBL height in hour *j*. Equation 134 135 (S1) is integrated from the surface to the top of boundary layer during daytime. 136 The daytime average O₃ production or loss rate integrated in the PBL is integrated to the 137 maximum PBL height, which is the mixing extent of O₃ in the PBL. It is computed using the 138 following equation, 139 $\overline{x} = \frac{\sum_{j} \sum_{i} x_{i}^{j} \cdot z_{i}^{j}}{\sum h_{\max}}$ 140 (S2) where *i* is the layer index, *j* is hour index, x_i^j is the hourly production or loss rate in layer *i* and hour 141 j, z_i^j is the hourly thickness of layer i in hour j, and h_{max} is the maximum hourly PBL height of the 142 143 day. Equation (S1) is integrated from the surface to the top of the maximum hourly PBL height 144 during daytime.

145 The daytime average OPE integrated in the PBL is computed similarly,
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$$OPE = \frac{\sum_{j} \sum_{i} P(O_3)_i^j \cdot z_i^j}{\sum \sum P(NO_i)^j \cdot z^j}$$
(S3)

 $\sum_{j} \sum_{i} P(NO_{z})_{i}^{j} \cdot Z_{i}^{j}$ 147 where *i* is the layer index, *j* is hour index, P_{i}^{j} is the hourly production rate of O₃ or NO_z in layer *i* 148 and hour *j*, and z_{i}^{j} is the hourly thickness of layer *i* in hour *j*. Equation (S1) is integrated from the

- surface to the top of the maximum hourly PBL height during daytime.
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Tables and figures

Species	Reactions		Mixing ratio (ppbv)	MGLY contribution (%)
Benzene	BENZENE + OH = 0.29 GLYX + Products	1.22E-12	2.2	1
Ethylbenzene	ETHYLBZ + OH = 0.184GLYX + 0.117MGLY + Products	7.00E-12	3.2	4
Toluene	TOLUENE + OH = 0.238 GLY + 0.151 MGLY +Products	5.58E-12	5.9	8
m-Xylene ¹	M-XYLENE + $OH = 0.38$ MGLY + Products	2.31E-11	2.2	31
o-Xylene	O-XYLENE + OH = 0.238 MGLY + 0.084GLYX + 0.185BACL + Products	1.36E-11	1.7	9
p-Xylene ¹	P-XYLENE + OH = 0.112 MGLY + 0.286GLYX + Products	1.43E-11	2.2	6
1,2,3- Trimethyl Benzene 1,2,4-	123TMB + OH = 0.072 MGLY + 0.18GLYX + 0.447BACL+ Products	3.27E-11	0.3	1
Trimethyl	124TMB + OH = 0.405 MGLY + 0.074 GLYX + 0.112 BACL+ Products	3.25E-11	0.4	8
Benzene 1,3,5-				
Trimethyl Benzene	135TMB + OH = 0.64 MGLY + Products	5.67E-11	0.5	32
ARO1 ²	ARO1 + OH = 0.218 GLYX + 0.138 MGLY + Products	6.15E-12	11.3	15
$ARO2^3$	ARO2 + OH = 0.116 GLYX + 0.286 MGLY + 0.104 BACL + Products	2.57E-11	7.3	85

197	Table S1. Reactions and rate constants for aromatics	s species with OH used in SAI	PRC-07 for deriving the lumped mechanism.
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¹ These two species were measured as one species mp-xylene, and we assume the same amount for each in this calculation. ² ARO1 includes benzene (\times 30%), ethyl-benzene, and toluene. ³ ARO2 includes all aromatics other than those included in ARO1.

Alkanes			Alkenes			
C2H6¹	Ethane ²	3.82	C2H4	ethene	3.6	
C3H8	Propane ³	3.73		Propene	1.0	
	isobutane	1.85		trans-2-butene	0.5	
	n-butane	2.14		1-butene	0.5	
	isopentane	3.50		Isobutene	0.3	
	n-pentane	1.22		cis-2-butene	0.4	
	2,2-dimthylbutane	0.07		3-methyl-1-butene	0.0	
	cyclopentane	0.14		1,3-Butadiene	0.3	
	2-methylpentane	0.69		1-pentene	0.5	
	3-methylpentane	0.50	PRPE	Isoprene	0.8	
	n-hexane	1.40		trans-2-pentene	0.2	
	2,4-dimethylpentane	0.06		cis-2-pentene	0.0	
	methylcyclopentane	0.47		2-methyl-2-butene	0.1	
ALK4	cyclohexane	0.18		Cyclopentene	0.0	
	2-methylhexane	0.21		4-methyl-1-pentene	0.0	
	2,3-dimethylpentane	0.09		2-methyl-1-pentene	0.0	
	3-methylhexane	0.25		trans-2-hexene	0.0	
	2,2,4-trimethylpentane	0.47		cis-2-hexene	0.0	
	n-heptane	3.47		alpha-pinene	0.1	
	methylcyclohexane	0.14		Aromatics		
	2,3,4-trimethylpentane	0.03		Benzene	2.1	
	2-methylheptane	0.09		Toluene	5.8	
	3-methylheptane	0.09	ARO1	Styrene	0.1	
	n-octane	0.91	_	Ethylbenzene	3.2	
	n-nonane	0.11		m,p-xylene	4.3	
	Alkyne			o-xylene	1.6	
C2H2	ethyne	5.67		Isopropylbenzene	0.0	
	-		ARO2	n-propylbenzene	0.0	
				1,3,5-trimethylbenzene	0.5	
				1,2,4-trimethylbenzene	0.3	
				1,2,3-trimethylbenzene	0.3	

202 **Table S2.** Average mixing ratios (ppbv) of explicit and lumped NMHCs in the model.

 ¹ Model species are listed to the left and measured species are listed to the right. Lumped model
 species include multiple measured species.

²Species listed in normal font were measured at 8:00-9:00 and 13:00-14:00 each day; the averages of these measurements are shown.

³Species listed in bold font were measured continuously each hour; 24-hour averages are shown.

Table S3. Daytime (6:00 - 18:00) average mixing ratios (ppbv) of formaldehyde, acetaldehyde, methyl glyoxal and glyoxal from observations and the model S0 and S2 (without aromatics); model values are sampled from the days with observed data.

	Formaldehyde	Acetaldehyde	Methylglyoxal	Glyoxal
Observed	12.4	7.0	0.9	3.1
SO	10.9	6.8	1.0	3.5
S2	4.9	4.1	0.2	0.3



Figure S1. Average diurnal profiles and breakdowns of O₃ production rates (ppbv h⁻¹) integrated in





Figure S2. Vertical profiles of daytime averaged vertical distribution of P(O₃) from the surface to 1 km for simulations S0 and S2.







Figure S4. Changes of PBL average O_3 production ($\Delta P(O_3)$) as a function of NOx, VOCs, and both for simulations S0 and S2.



Figure S5. Hourly $\Delta P(O_3)$ integrated in the PBL due to NOx changes under S0 and S3a.